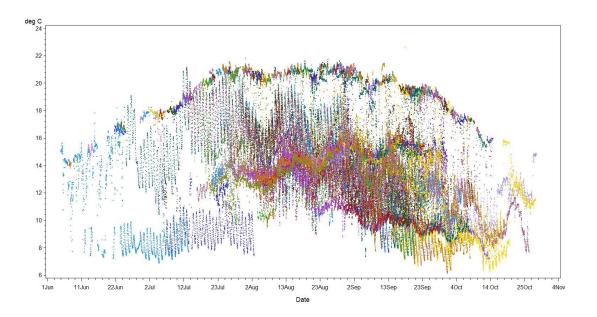
ASSEMBLY AND ANALYSIS OF RADIOTELEMETRY AND TEMPERATURE LOGGER DATA FROM ADULT CHINOOK SALMON AND STEELHEAD MIGRATING THROUGH THE COLUMBIA RIVER BASIN

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1.0 Introduction

1.1 Background: cold water refuge use by adult salmon and steelhead

Summertime water temperatures in the lower Columbia River have steadily increased over the last several decades. Annual peak temperatures exceeded 21 °C in most recent years and have been as high as 24 °C. The warmest period typically occurs from late July to early September, coincident with late-migrating adult summer Chinook salmon (*Oncorhynchus tshawytscha*) and sockeye salmon (*O. nerka*) and with substantial portions of the adult fall Chinook salmon and summer steelhead (*O. mykiss*) runs. Water temperatures in the 19-22 °C range, like those that routinely occur in the Columbia River and Snake River main stems, are a significant management concern for adult migrants. Portions of many populations can experience thermal conditions that have been associated with behavioral changes and a variety of sub-lethal effects on physiology, disease susceptibility, reproductive development, gamete quality, survival, and fitness.

Many adult salmon and steelhead behaviorally thermoregulate by using cold water refuge (CWR) sites when Columbia and Snake River water temperatures are high. Most of the CWR sites temporarily used by adult migrants are located where cold- or cool-water tributaries enter Columbia and Snake River reservoirs. Extensive adult use of these CWR sites was identified during a series of radiotelemetry studies conducted by the University of Idaho (UI) and NOAA-Fisheries. Basic descriptions of some of the most-used CWR sites, the timing of CWR use by adult migrants, and ambient temperature thresholds associated with CWR use have been described previously (Keefer et al. 2004; Goniea et al. 2006; High et al. 2006; Keefer et al. 2009, 2011; Keefer and Caudill 2016).

In 2000 and 2002, a subset of the radio transmitters used in adult Chinook salmon and steelhead also had onboard sensors that recorded fish depth and internal body temperature, a combination referred to as radio data storage tags (RDST). The RDST study component allowed for highly-detailed, spatially- and temporally-explicit reconstructions of the thermal experiences of migrants. Portions of these histories were reported previously by Clabough et al. (2008) and Caudill et al. (2013).

1.2 Use of RDST data for HexSim modeling

The RDST data were identified as a potentially useful source for parameterizing and validating the individual-based HexSim model (Schumaker 2016) that is currently being developed by EPA for adult salmon and steelhead. In this application, the HexSim model is intended to reproduce and simulate the thermal experiences of adult migrants as they move through the main stem Columbia River migration corridor from Bonneville Dam to the Columbia River–Snake River confluence, including their use of CWR sites. This report describes the excerpting, processing, and compilation of the existing RDST data and provides quality assurance documentation. The report also includes complete thermal histories, a variety of RDST data summaries, and classifications of tagged fish behaviors in relation to CWR use. The data summaries were developed in consultation with EPA, specifically to inform the HexSim modeling effort.

2.0 Methods and Data Quality Assurance

2.1 Data collection

The RDST data were collected in 2000 and 2002 during studies of adult summer- and fall-run Chinook salmon and summer-run steelhead collected at Bonneville Dam, released downstream from the dam, and then monitored as they migrated upstream past dams, through reservoirs, and into tributaries. The transmitters (Model LTD_100, Lotek Wireless, Inc.) were programmed to log fish body temperature every one minute and pressure (i.e., fish depth) every 5 seconds. These settings allowed the transmitters to store approximately 40 days of data and produced data files of ~600,000 combined radiotelemetry, temperature, and depth records per fish. The manufacturer-reported specifications for the transmitters included temperature accuracy of ± 0.15 °C at ambient temperatures of 0-20 °C and ± 0.10 °C at temperatures of 20-35 °C; resolution was reported to be 0.02 °C.

Data stored on the transmitters were not available unless the tags were recovered, so salmon and steelhead with relatively high recapture probability were preferentially tagged. Samples were predominated by hatchery-origin fish and by fish from populations whose migrations passed adult trapping facilities, especially the Lower Granite Dam trap in the lower Snake River. Adult fish origin at the time of RDST-tagging was assessed using the presence of hatchery fin clips and/or the presence of a passive integrated transponder (PIT) tag that was implanted in juvenile fish (see Keefer et al. 2009 for details). Majorities of the tagged samples in 2002 (but not 2000) had juvenile PIT tags, with many individuals from the Snake River basin (see Results).

The sampling broadly reflected seasonal run timing, but the prerogative to recover the RDST data outweighed our effort to tag random subsamples of the runs at large. Juvenile PIT-tagging projects in the late 1990s and early 2000s were disproportionately focused on interior populations (i.e., from upper Columbia River and Snake River populations) and the RDST-tagged fish reflect these biases. Furthermore, the recovered RDST data were primarily from 'successful' migrants that reached adult trapping facilities. Fish that died naturally during migration (including due to high temperature exposure) and those that were harvested were almost certainly under-represented in the dataset. These sampling biases should be considered while interpreting results.

2.2 Monitoring of radio-tagged fish

Radio-tagged salmon and steelhead were monitored in Columbia and Snake River dam tailraces, outside and inside dam fishways, in reservoirs, and in tributaries to assess a diverse mix of study objectives (Figures 1-3). The raw telemetry files were filtered (i.e., 'coded') by UI biologists using a set of rule-based decisions, and the coded subsets were then vetted by experienced personnel. Radiotelemetry detection data were integrated with the RDST data so that each temperature history has a spatial element; the spatial resolution varies among river reaches as a function of the number and placement of antennas. The aerial (Yagi) radio antennas used to monitor fish passage through dam tailraces, reservoirs, and near CWR sites were tuned to detect

fish within approximately 0.5 km. However, actual detection ranges varied as a result of site configuration, fish depth (radio signals attenuate with depth), and transmitter signal strength.

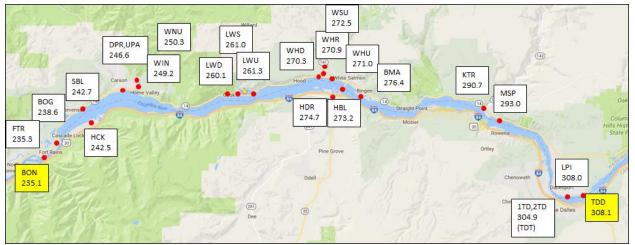


Figure 1. Map of the Columbia River from Bonneville Dam (BON) to The Dalles Dam (TDD) showing locations of aerial antennas (•) and radio receiver names and river kilometer (text boxes). Extensive underwater (coaxial cable) antenna arrays at dams are not shown.



Figure 2. Map of the Columbia River from The Dalles Dam (TDD) to John Day Dam (JDD) showing locations of aerial antennas (•) and radio receiver names and river kilometer (text boxes). Extensive underwater (coaxial cable) antenna arrays at dams are not shown.

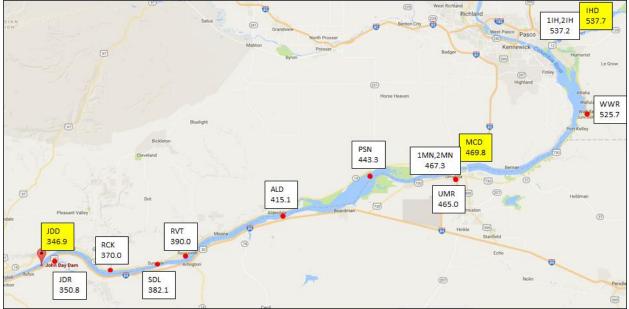


Figure 3. Map of the Columbia and lower Snake rivers from John Day Dam (JDD) past McNary Dam (MCD) to Ice Harbor Dam (IHD) showing locations of aerial antennas (•) and radio receiver names and river kilometer (text boxes). Extensive underwater (coaxial cable) antenna arrays at dams are not shown.

2.3 Recovery and processing of RDST data

Data from recovered RDST tags were downloaded and processed using the manufacturerrecommended methods. Quality control screens included tests for temperature values outside expected ranges and cross-checks with the radiotelemetry data for appropriate date-time stamps. To better match the HexSim model needs, we reduced the RDST temperature and depth data to 30-minute time steps (i.e., to 48 records per day). The radiotelemetry datasets were also substantially reduced so that only first detection records for each block of detections for each fish were included from aerial antenna sites and only first tailrace, first fishway approach, and top-ofladder exit records were included from antenna arrays at dams. The temperature, depth and radiotelemetry data were then merged to assemble combined detection+temperature histories for each fish.

An additional post-processing step was to assign each RDST temperature record to a specific river reach. Reaches included individual Columbia and Snake River reservoirs (last record at downstream dam to first approach record at upstream dam) and individual dams (first fishway approach record to pass a ladder), or the Hanford Reach of the Columbia River. Two additional reaches were the Yakima River and the extended Columbia River reach that included multiple dams and reservoirs upstream from Priest Rapids Dam (see Figure 6) where there was limited radiotelemetry monitoring effort.

2.4 Classifying cold water refuge (CWR) use by RDST-tagged fish

Adult salmon and steelhead have highly variable behaviors as they migrate through Columbia River reservoirs and past dams, and use of cold water refuge (CWR) sites adds additional spatial and temporal variation. EPA requested a simple classification system for CWR use to inform the HexSim modeling effort. The classification used in this report includes three primary criteria: (1) an individual fish temperature threshold of at least 2 °C cooler than ambient river water as measured at nearby dam water quality monitoring (WQM) sites; (2) a cumulative individual exposure duration of at least 4 h at or below the 2 °C CWR temperature threshold; and (3) an indication of whether one or more CWR sites were used based on spatial telemetry records (Figure 4).

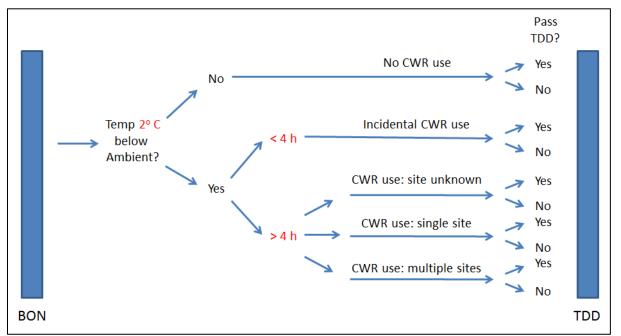


Figure 4. Classification tree for generalized use and non-use of cold water refuge (CWR) sites in a reservoir reach. Example is from Bonneville Dam (BON) to The Dalles Dam (TDD). In this schema, fish temperature (Temp) ≥ 2 °C cooler than ambient reservoir temperature was the evidential threshold for CWR use. 'Incidental use' was defined as < 4 h of cumulative fish temperature below ambient.

We included a CWR 'site unknown' classification to accommodate histories where fish location was ambiguous or unknown. Ambiguous areas included tributary plumes which were often incompletely monitored with the radiotelemetry antennas. In some cases fish were also detected on antennas that were near, but not in, a tributary. Location uncertainty was especially high in the thermally complex stretch of habitats in the Bonneville reservoir between the Wind and Klickitat rivers that included the closely-spaced Little White Salmon and White Salmon river confluences with the Columbia River. Other 'unknown' sites included unmonitored tributaries (i.e., several smaller creeks) and potential non-tributary sources of cold water such as springs, seeps or sites like the outfall from Irrigon Hatchery (adjacent to John Day reservoir).

Two complicating factors in the CWR use classification were whether or not an individual passed through a given river reach and whether RDST data were collected throughout the reach. Failure to pass a reach due to harvest or natal tributary entry adds a layer of uncertainty to the proportions of fish classified in each CWR use category. Incomplete RDST data due to storage capacity limitations similarly contributes to classification uncertainty. These factors are noted, where relevant, in the Results.

3.0 Data Summaries

3.1 Columbia River water temperature: 2000 and 2002

Columbia River temperatures were near average or slightly warmer than average during the summers (June-August) of 2000 and 2002 (Figure 5). In contrast, temperatures were well below average during September and October in both years. These patterns are important for interpreting the RDST data and fish use of CWR sites. The cool fall temperatures, in particular, suggest that late-run steelhead and many fall Chinook salmon may have been less likely to use CWR sites in 2000 and 2002 than in average or warm temperature years.

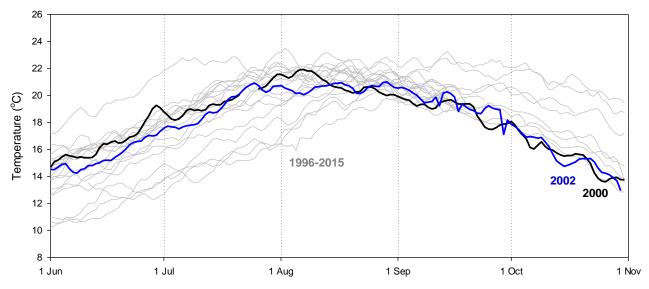


Figure 5. Mean daily Columbia River water temperature recorded at the Bonneville Dam water quality monitoring (WQM) site, 1996-2015. Note that some years have incomplete data in the fall. Source: Columbia River DART (www.cbr.washington.edu/dart)

The WQM monitoring sites collected temperature at Bonneville and McNary dams throughout the summer and fall RDST study period (June-November) in both years. Data were collected at The Dalles and John Day dams for shorter periods each year, with end dates in mid-September through early November (Table 1). Mean daily water temperatures were highly correlated (r > 0.96) among pairs of dams within each year for the date ranges when data were available.

		The Dalles ¹	John Day ²	McNary
2000	Bonneville	0.994	0.980	0.994
	The Dalles	-	0.989	0.976
	John Day	-	-	0.969
		The Dalles ³	John Day ⁴	McNary
2002	Bonneville	0.987	0.987	0.986
	The Dalles	-	0.988	0.983
	John Day	-	-	0.983

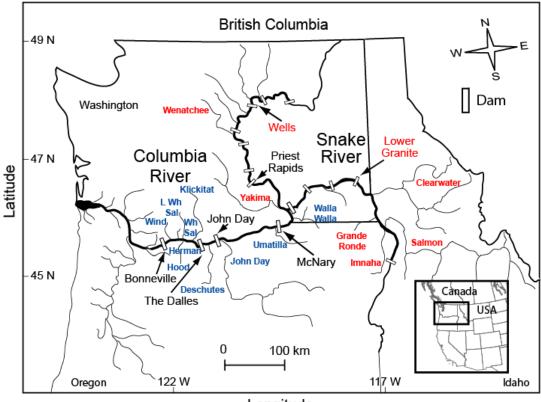
Table 1. Correlation matrix of WQM daily average water temperature data for available data between 1 June and 30 November. Source: Columbia River DART (www.cbr.washington.edu/dart)

¹ 20 September end date; ² 21 September end date; ³ 6 November end date; ⁴ 27 September end date

3.2 Origin of recovered RDST-tagged fish

In total, usable DST temperature data were recovered for 73 summer Chinook salmon, 58 fall Chinook salmon, and 284 steelhead (Table 2). In 2000, 85-98% of the RDST-tagged fish had no juvenile PIT tag (i.e., their natal origin was unknown); the largest group of known-origin fish in 2000 were six summer Chinook salmon tagged at hatcheries in the upper Columbia River basin. In contrast, majorities of the tagged fish in 2002 did have juvenile PIT tags, including 67% of summer Chinook salmon, 77% of fall Chinook salmon, and 90% of steelhead. Origin sites for the 2002 samples were predominantly in the Snake River basin, reflecting large juvenile PIT-tag programs there in the late 1990s and early 2000s (Table 2, Figure 6).

Table 2. Locations where RDST-tagge	Table 2. Locations where RDST-tagged adults were PIT tagged as juveniles.									
	Summer	Fall Chinook			Steelhead					
PIT-site	2000	2002	2000	2002	2000	2002				
Total recovered with temperature data	40	33	36	22	116	168				
Juvenile PIT-tag site										
Snake River basin										
Lower Granite Dam		10		1		114				
Snake River		1		6		2				
Clearwater River			1	10	1	16				
Grande Ronde River		5				9				
Salmon River		6			1	7				
Imnaha River						3				
Upper Columbia River										
East Bank Hatchery	4									
Wells Hatchery	2					1				
No juvenile PIT tag site	34	11	35	5	114	16				



Longitude

Figure 6. Map showing locations of dams, cold water refuge (CWR) tributaries (blue text) and locations where RDST-tagged adults were PIT-tagged as juveniles (red text).

3.3 Representativeness of recovered RDST-tagged fish

In 2000, the recovered RDST-tagged fish were tagged throughout their respective run-timing distributions at Bonneville Dam (Figure 7). However, median tag dates were later than medians based on run-specific counts by 6 d (summer and fall Chinook salmon) and 9 days (steelhead). Tagged fish were also under-represented during the tails of each run timing distribution, an artifact of attempting to tag approximately in proportion to the runs at large based on timing in previous years. In 2002, the recovered fish were similarly tagged throughout their respective run-timing distributions (Figure 8). Median tag dates were again later than medians based on run-specific counts, but by just 2 d (summer and fall Chinook salmon) and 8 days (steelhead).

As noted previously, juvenile PIT tags identified natal river basin for very few of the RDSTtagged fish in 2000 but for majorities of the 2002-tagged fish. Selection for adults with juvenile PIT tags improved the likelihood of RDST recovery, but had a mix of known and uncertain effects on sample representativeness. The previously PIT-tagged fish were more likely to be of non-hatchery origin, for example, primarily due to collection of wild smolts for PIT tagging at Lower Granite Dam. The percentage of each sample with fin clips indicating hatchery origin were: 100% (2000 summer Chinook), 58% (2002 summer Chinook), 100% (2000 fall Chinook), 10% (2002 fall Chinook), 98% (2000 steelhead), and 21% (2002 steelhead). We think it is likely that Snake River populations were over-sampled in 2002 relative to the runs at large, especially for fall Chinook salmon, and this should be considered when interpreting results. Furthermore, all recovered RDST data in both years were from fish that were either harvested or that returned to traps or hatcheries, and it is possible that some recovery bias affected the representativeness of the data. For example, wild fish and fish that died prematurely for non-harvest reasons were likely under-represented.

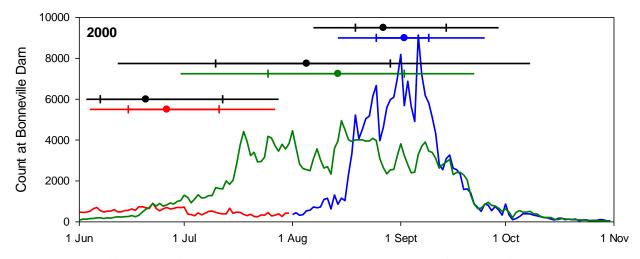


Figure 7. Daily adult Chinook salmon and steelhead counts at Bonneville Dam (lines: red = summer Chinook, blue = fall Chinook, green = steelhead) and tag dates of RDSTs recovered in 2000. Black whisker plots show 5th, 25th, 50th, 75th, and 95th percentiles of RDST tagging dates (recovered transmitters only) and colored whisker plots are for adult counts at Bonneville Dam which correspond with solid lines (Source: Columbia River DART).

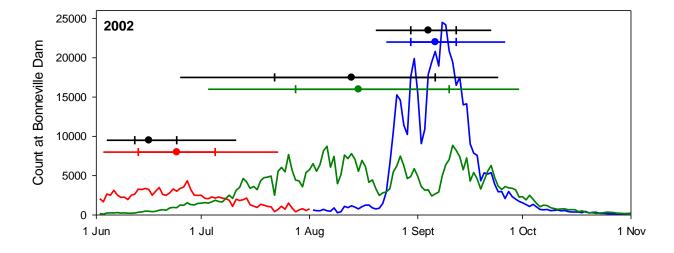


Figure 8. Daily adult Chinook salmon and steelhead counts at Bonneville Dam (lines: red = summer Chinook, blue = fall Chinook, green = steelhead) and tag dates of RDSTs recovered in 2002. Black whisker plots show 5th, 25th, 50th, 75th, and 95th percentiles of RDST tagging dates (recovered transmitters only) and colored whisker plots are for adult counts at Bonneville Dam which correspond with solid lines (Source: Columbia River DART).

3.4 Last detection locations of recovered RDST-tagged fish

Summer Chinook salmon were last detected at sites throughout the Columbia River basin (Table 3). In both years, about 10% were last in reservoirs, at dams, or in tributaries in the lower Columbia basin downstream from the Snake River confluence. Another 35% (2000) and 73% (2002) were in the Snake River basin and 55% (2000) and 18% (2002) were at upper Columbia River sites above the Snake River confluence.

Last detection locations for fall Chinook salmon in 2000 included 53% in the lower Columbia River, 31% in the Snake River, and 8% at upper Columbia River sites above the Snake River (Table 3). The distribution in 2002 included 32% at lower Columbia River sites, 64% in the Snake River, and 5% at upper Columbia River sites. These distributions are consistent with over-sampling of Snake River fish.

Last detection locations for steelhead in 2000 included 49% in the lower Columbia River, 44% in the Snake River, and 3% at upper Columbia River sites (Table 3). The distribution in 2002 included 20% at lower Columbia River sites, 79% in the Snake River, and 1% at upper Columbia River sites. Difference in distributions between years reflected the selection for previously-PIT tagged fish in 2002.

3.5 Thermal histories of RDST-tagged fish

3.5.1 Dataset summary

In total, 415 RDST transmitters were recovered with more than 540,000 temperature records excerpted at 30-minute intervals (Table 4). By run, the data represented 1,759 d (summer Chinook salmon), 1,062 d (fall Chinook salmon), and 8,505 d (steelhead) of body temperatures during migration.

The spatial distribution of where temperature data were collected varied among runs, reflecting differences in run timing, natal origin, and migration behaviors. Many fish were in the lower Columbia River study area for longer than the 40-d data storage capacity for the RDSTs and this resulted in disproportionately higher data collection rates in the lower reaches (i.e., Bonneville and The Dalles pools), especially for steelhead.

From ~10-22% of the temperature records from each run-year were from sites at or downstream from Bonneville Dam, collected as fish moved upstream from the release sites ~10 km below the dam (Table 4). The percent of records collected in the Bonneville pool reach were 7-8% (summer Chinook salmon), 11-28% (fall Chinook salmon), and 46-50% (steelhead), reflecting the among-run differences in CWR use in this reach. Percentages in The Dalles pool reach were ~5% (summer Chinook salmon), 6-8% (fall Chinook salmon), and 11-19% (steelhead), similarly reflecting extended use of the Deschutes River CWR by steelhead relative to Chinook salmon.

	on adult recovery location in relation to juvenile PTT tag location. Summer Fall Chinook Steelhead										
RDST recovery site	2000	2002	2000	2002	2000	2002					
Total recovered with data	40	33	36	2002	116	168					
Total recovered with data	40	55	50	22	110	100					
Lower Columbia River basin											
Sandy River					1						
Below Bonneville Dam				1	-	2					
Bonneville pool	3	2	1	2	25	16					
Herman / Eagle Creek	C	-	-	-	1	1					
Little White Salmon River			1		4	-					
White Salmon River			-		2						
Klickitat River			4		4	4					
Hood River					2						
The Dalles pool	1	1	4	1	4	6					
Deschutes River			1		2	-					
John Day pool			7	3	8	5					
John Day River					1						
Umatilla River			1								
McNary pool					6						
Walla Walla River					1						
Snake River basin											
Ice Harbor pool			1		1	1					
Lower Monumental pool					1						
Lyons Ferry Hatchery			2								
Little Goose pool			3	1		1					
Lower Granite dam / trap		9	2	1	6	63					
Lower Granite pool	2		_	_	3	1					
Snake River > L. Gr. pool			3	3	7	14					
Clearwater River	1	_	3	9	18	17					
Grande Ronde River	1	6				5					
Salmon River	8	8			15	28					
Imnaha River	2	1				2					
Upper Columbia River											
Yakima River			1								
Hanford Reach			2								
Ringold Hatchery			<i>L</i>	1	2						
Wanapum pool				1	1						
Wenatchee River	3				1						
East Bank Hatchery	5				1						
Rocky Reach pool	3	2									
Wells trap	8	-				1					
Wells pool	3	3				1					
Methow River	1	2				-					
Okanogan River	4	1									
<i>o o o o o o o o o o</i>		-									

Table 3. Locations where RDST-tagged fish were last detected. See Appendix A for additional details on adult recovery location in relation to juvenile PIT tag location.

lower Snake and upper	Columbia reache	es include dat	ta from dams, res	ervoirs, and C	CWR sites.	
	Summer Chine	ook	Fall Chinook		Steelhead	
	2000	2002	2000	2002	2000	2002
Number of fish	40	33	36	22	116	168
Number of DST temp	erature records	(all sites)				
30-minute	54,254	30,201	34,653	16,325	151,908	256,313
Days	1,130	629	722	340	3,165	5,340
Number of DST temp	erature records	per reach (i	ncluding CWRs)		
Below Bonneville ¹	5,552	2,928	5.524	3,646	15,901	26,172
Bonneville pool	4,492	1,964	9,554	1,751	75,197	116,693
The Dalles pool	2,633	1,593	2,178	1,271	16,192	47,898
John Day pool	3,177	4,071	3,723	1,603	8,354	12,228
McNary pool	2,376	1,700	1,555	1,112	9,098	13,558
Yakima River	-	-	1,205	-	-	-
Hanford Reach	2,023	657	602	1,173	7,536	3,960
Lower Snake (all)	3,473	4,357	5,100	3,121	7,389	14,947
Upper Columbia (all)	24,286	5,391	-	-	3,135	1,415
Percent of DST tempe	erature records	per reach (ir	cluding CWRs)			
Below Bonneville ¹	10.2%	9.7%	15.9%	22.3%	10.5%	10.2%
Bonneville pool	8.3%	6.5%	27.6%	10.7%	49.5%	45.5%
The Dalles pool	4.9%	5.3%	6.3%	7.8%	10.7%	18.7%
John Day pool	5.9%	13.5%	10.7%	9.8%	5.5%	4.8%
McNary pool	4.4%	5.6%	4.5%	6.8%	6.0%	5.3%
Yakima River	-	-	3.5%	-	-	-
Hanford Reach	3.7%	2.2%	1.7%	7.2%	5.0%	1.5%
Lower Snake (all)	6.4%	14.4%	14.7%	19.1%	4.9%	5.8%
Upper Columbia (all)	44.8%	17.9%	-	-	2.1%	0.6%

Table 4. Summary of recovered RDST data in total and by migration reach. Reaches in the lower Columbia River (the primary study area) include records in pools and CWR sites, but not at dams; the lower Snake and upper Columbia reaches include data from dams, reservoirs, and CWR sites.

¹ includes Bonneville Dam

Most salmon and steelhead moved relatively quickly through the John Day pool and McNary pool reaches, where there were few CWR sites. The percentages of records collected in these reaches were therefore more similar among runs, ranging from 4-14% (summer Chinook salmon), 5-11% (fall Chinook salmon), and 5-6% (steelhead) per reach (Table 4). The remaining temperature data were from the lower Snake River, Hanford Reach, Yakima River, and upper Columbia River.

Annotated salmon and steelhead temperature histories, with use of CWR sites identified, are included in Appendix B for 9 summer Chinook salmon, 7 fall Chinook salmon, and 11 steelhead with a variety of migration behaviors.

Temperature histories for all 415 fish, color coded for river reach but not CWR use, are included in Appendix C (2000) and Appendix D (2002).

3.5.2 Bonneville reservoir reach: Chinook salmon

In the two years, 73 summer Chinook salmon had RDST data in the Bonneville pool reach and 69 (95%) passed The Dalles Dam (Table 5, Figures 9 and 10). A large majority (96% of 73) of the salmon had no temperature history evidence of CWR use. One had incidental use < 4 h, one used a single CWR, and one used multiple CWR sites and each of these three fish were among the 10 that encountered water temperature ≥ 20 °C. Likely use sites, based on telemetry detections, included the Little White Salmon, White Salmon, Hood, and Klickitat rivers.

A total of 58 fall Chinook salmon had RDST data and 50 (86%) of these passed The Dalles Dam. A majority (57%, n = 33) had no evidence of CWR use, 10% (n = 6) had incidental use, 9% (n = 5) used unknown CWR sites, and 24% (n = 14) were detected at one or more CWR sites (Table 5, Figures 9 and 10). CWR use rates were somewhat higher for fish without fin clips and for those that encountered water temperature ≥ 20 °C. Several fall Chinook salmon were detected in the Klickitat, Little White Salmon, and White Salmon rivers. The unknown sites were likely tributary plumes but may have included unmonitored sites.

from th	from the Bonneville Dam WQM site were used as the ambient temperature reference.											
		Passed		Cold water rea	fuge (CWR) ເ	use						
Year	Run	The Dalles	п	None	Incidental	Unk CWR	1 CWR	>1 CWR				
2000	Summer	Yes	38	36 (95%)	1 (3%)		1 (3%)					
		No	2	1 (50%)				1 (50%)				
2002	Summer	Yes	31	31 (100%								
		No	2	2 (100%)								
Both	Summer-W	Yes+No	14	14 (100%)								
	Summer-H	Yes+No	59	56 (95%)	1 (2%)		1 (2%)	1 (2%)				
Both	$T \ge 20^{\circ}C$	Yes+No	10	7 (70%)	1 (10%)		1 (10%)	1 (10%)				
Both	$T < 20^{\circ}C$	Yes+No	63	63 (100%)								
• • • • •	N 11	**			1 (1 2 2 ()	2 (100)	1 (1 2 2 ()	2 (1 2 2 ()				
2000	Fall	Yes	31	17 (55%)	4 (13%)	3 (10%)	4 (13%)	3 (10%)				
		No	5				2 (40%)	3 (60%)				
2002	Fall	Yes	19	15 (79%)	2 (11%)	1 (5%)	1 (5%)					
		No	3	1 (33%)		1 (33%)	1 (33%)					
Both	Fall-W	Yes+No	20	14 (70%)	2 (10%)	2 (10%)	2 (10%)					
	Fall-H	Yes+No	38	19 (50%)	4 (11%)	3 (8%)	6 (16%)	6 (16%)				
Both	$T \ge 20^{\circ}C$	Yes+No	38	20 (53%)	5 (13%)	4 (11%)	6 (16%)	3 (8%)				
Both	$T < 20^{\circ}C$	Yes+No	20	13 (65%)	1 (5%)	1 (5%)	2 (10%)	3 (15%)				

Table 5. Numbers of summer and fall Chinook salmon with RDST temperature data collected in the Bonneville pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the Bonneville Dam WQM site were used as the ambient temperature reference.

Notably, four fall Chinook salmon were last detected in the Klickitat River in 2000 (see Table 3 and Appendix A), including three that were reported harvested. The natal origins of these fish were unknown and therefore they may have originated from the Klickitat River (i.e., the behavior was not CWR use). Alternately, these fish may be examples of harvest of non-local thermoregulating fish.

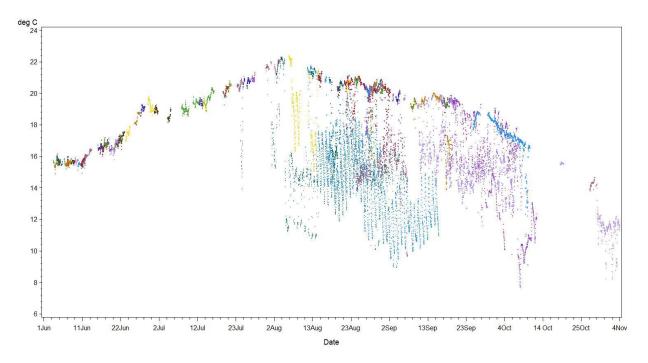


Figure 9. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 39) and fall Chinook salmon (n = 36) in Bonneville reservoir in 2000. Different colors represent individual fish.

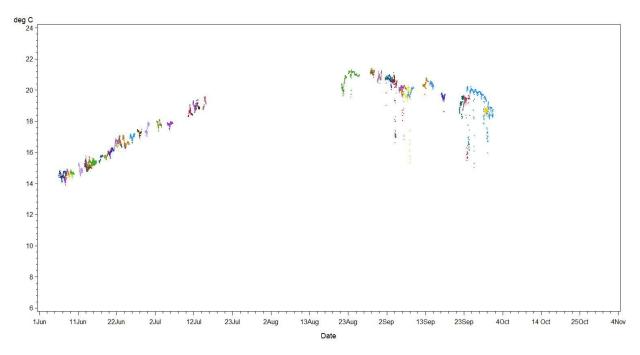


Figure 10. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 33) and fall Chinook salmon (n = 22) in Bonneville reservoir in 2002. Different colors represent individual fish.

3.5.3 Bonneville reservoir reach: Steelhead

A total of 281 steelhead had RDST data in the reach and 223 (78%) of these passed The Dalles Dam (Table 5, Figures 11 and 12). In contrast to both Chinook salmon runs, a minority (30%, n = 83) of steelhead had no evidence of CWR use. The non-use group included many steelhead that migrated relatively early or late in the season as well as a subset (n = 25) that did not pass The Dalles Dam, mostly due to harvest in the reservoir or in CWR sites (see Table 3 and Appendix A). CWR behaviors were broadly similar for wild and hatchery-origin steelhead. Those that encountered water temperature ≥ 20 °C were far more likely to use a CWR than those that experienced cooler reservoir temperatures.

Incidental CWR use was detected for 9% (n = 26) of steelhead. Steelhead that were classified as using CWR included 33% (n = 92) detected at a single CWR site, 23% (n = 64) detected at multiple CWR sites, and 6% (n = 16) used unknown sites (Table 6). Some steelhead were detected at all of the known CWR locations in the Bonneville pool reach, including Herman Creek and the Wind, Little White Salmon, White Salmon, Hood and Klickitat rivers. The Little White Salmon River was the most-used CWR site in the reach.

Several steelhead were last recorded in Bonneville pool tributaries, especially in the Klickitat and Little White Salmon rivers (Table 3). A majority of these were reported harvested, but natal origins were mostly unknown.

Table 6. Numbers of steelhead with RDST temperature data collected in the Bonneville pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the Bonneville Dam WQM site were used as the ambient temperature reference.

		Passed		Cold water				
Year	Run	The Dalles	п	None	Incidental	Unk CWR	1 CWR	>1 CWR
2000	Steelhead	Yes	78	24 (31%)	9 (12%)	1 (1%)	23 (29%)	21 (27%)
		No	37	15 (41%)	1 (3%)		13 (35%)	8 (22%)
2002	Steelhead	Yes	145	34 (23%)	15 (10%)	13 (9%)	54 (37%)	29 (20%)
		No	21	10 (48%)	1 (5%)	2 (10%)	2 (10%)	6 (29%)
Both	Steelhead-W	Yes+No	133	36 (27%)	15 (11%)	11 (8%)	46 (35%)	25 (19%)
	Steelhead-H	Yes+No	148	47 (32%)	11 (7%)	5 (3%)	46 (31%)	39 (26%)
Both	$T \ge 20^{\circ}C$	Yes+No	175	23 (13%)	17 (10%)	8 (5%)	74 (42%)	53 (30%)
Both	$T < 20^{\circ}C$	Yes+No	106	60 (57%)	9 (8%)	8 (8%)	18 (17%)	11 (10%)

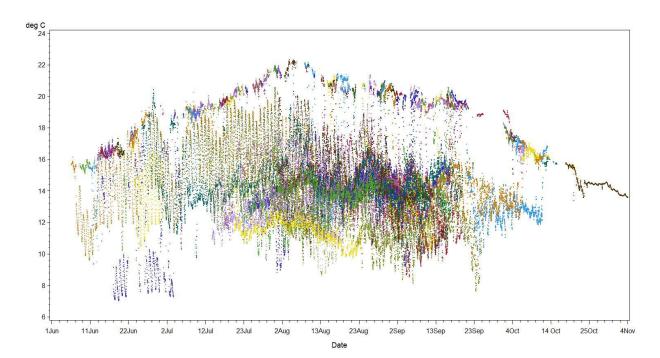


Figure 11. Thermal histories (30-min) of RDST-tagged steelhead (n = 115) in Bonneville reservoir in 2000.

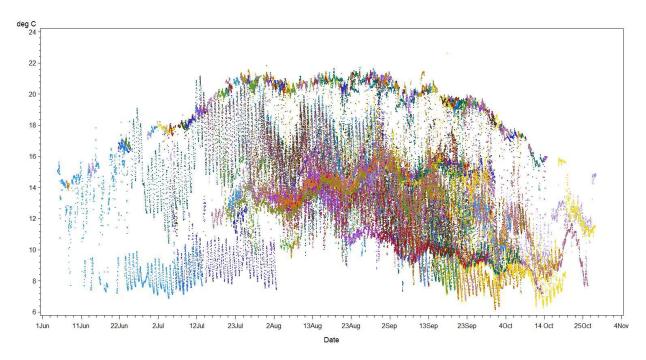


Figure 12. Thermal histories (30-min) of RDST-tagged steelhead (n = 166) in Bonneville reservoir in 2002.

3.5.4 The Dalles reservoir reach: Chinook salmon

A total of 69 summer Chinook salmon had RDST data and 64 (93%) of these passed John Day Dam (Table 7). A large majority (96%, n = 66) had no evidence of CWR use in the Deschutes River, the only known CWR in the reach (Figures 13 and 14). Two salmon (3% of 69) used the Deschutes River or its plume and one (1%) had incidental use < 4 h. The CWR use occurred when reservoir temperatures were ≥ 20 °C.

A total of 47 fall Chinook salmon had RDST data and 42 (89%) of these passed John Day Dam (Table 7). A majority (74%, n = 35) had no evidence of CWR use in the Deschutes River (Figures 13 and 14). Eleven salmon (23% of 47) used the Deschutes River or plume and three (6%) had incidental use < 4 h. A single unknown-origin fall Chinook salmon was last recorded in the Deschutes River (Table 3). Patterns were generally similar for wild and hatchery salmon and for those that did or did not encounter reservoir temperatures were ≥ 20 °C.

Table 7. Numbers of summer and fall Chinook salmon with RDST temperature data collected in The Dalles pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the Bonneville Dam WQM site were used as the ambient temperature reference because The Dalles WQM data were not available through the fall.

		Passed		Cold water refuge (CWR) use					
Year	Run	John Day	n	None	Incidental	Unk CWR	1 CWR	>1 CWR	
2000	Summer	Yes	37	34 (92%)	1 (3%)		2 (5%)		
		No	1	1 (100%)					
2002	Summer	Yes	30	30 (100%					
		No	1	1 (100%)					
Both	Summer-W	Yes+No	13	13 (100%)					
	Summer-H	Yes+No	56	53 (95%)	1 (2%)		2 (4%)		
Both	$T \ge 20^{\circ}C$	Yes+No	6	4 (67%)			2 (33%)		
Both	$T < 20^{\circ}C$	Yes+No	62	61 (98%)	1 (2%)				
2000	Fall	Yes	24	19 (79%)	1 (4%)		5 (21%)		
		No	4	3 (75%)	~ /		1 (25%)		
		n/a^1	3						
2002	Fall	Yes	18	12 (67%)	2 (11%)		4 (22%)		
		No	1				1(100%)		
Both	Fall-W	Yes+No	17	11 (65%)	2 (12%)		4 (24%)		
	Fall-H	Yes+No	30	23 (77%)	1 (3%)		6 (20%)		
Both	$T \ge 20^{\circ}C$	Yes+No	30	23 (77%)	1 (3%)		6 (20%)		
Both	$T < 20^{\circ}C$	Yes+No	17	11 (65%)	2 (12%)		4 (24%)		

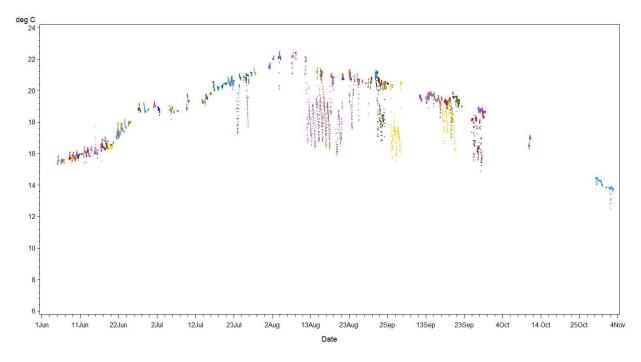


Figure 13. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 37) and fall Chinook salmon (n = 28) in The Dalles reservoir in 2000.

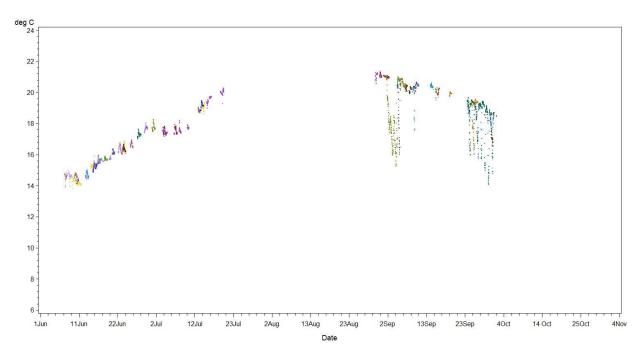


Figure 14. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 31) and fall Chinook salmon (n = 19) in The Dalles reservoir in 2002.

3.5.5 The Dalles reservoir reach: Steelhead

A total of 181 steelhead had RDST data in the reach and 170 (94%) of these passed John Day Dam (Table 8). Slightly more than half of the steelhead (51%, n = 93) met the CWR use criteria and another 9% (n = 17) had incidental use (Figures 15 and 16). Two unknown-origin steelhead were last recorded in the Deschutes River (Table 3 and Appendix A). Higher CWR use by wild steelhead was presumably related to the higher incidence of wild fish in the 2002 sample, when more fish used CWR generally. CWR use was higher across years among steelhead that encountered reservoir temperatures ≥ 20 °C.

Table 8. Numbers of steelhead with RDST temperature data collected in The Dalles pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the Bonneville Dam WQM site were used as the ambient temperature reference because The Dalles WQM data were not available through the fall.

		Passed		Cold water re	efuge (CWR)	use		
Year	Run	John Day	п	None	Incidental	Unk CWR	1 CWR	>1 CWR
2000	Steelhead	Yes	59	32 (54%)	6 (10%)		21 (36%)	
		No	6	3 (50%)			3 (50%)	
		n/a^1	13					
2002	Steelhead	Yes	111	33 (30%)	11 (10%)		67 (60%)	
		No	5	3 (60%)			2 (40%)	
		n/a ¹	29					
Both	Steelhead-W	Yes+No	102	30 (29%)	9 (9%)		63 (62%)	
	Steelhead-H	Yes+No	79	41 (52%)	8 (10%)		30 (38%)	
Both	$T \ge 20^{\circ}C$	Yes+No	89	27 (30%)	9 (10%)		53 (60%)	
Both	$T < 20^{\circ}C$	Yes+No	92	44 (48%)	8 (9%)		40 (43%)	

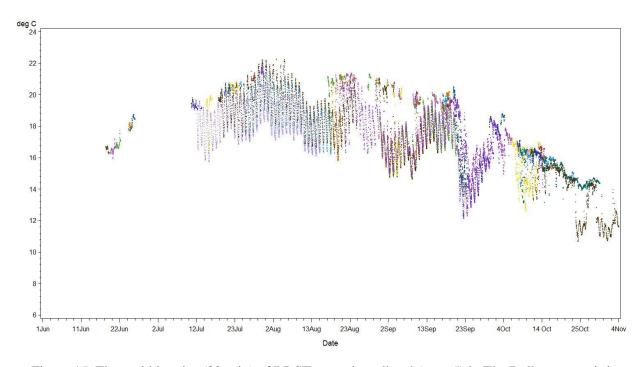


Figure 15. Thermal histories (30-min) of RDST-tagged steelhead (n = 65) in The Dalles reservoir in 2002.

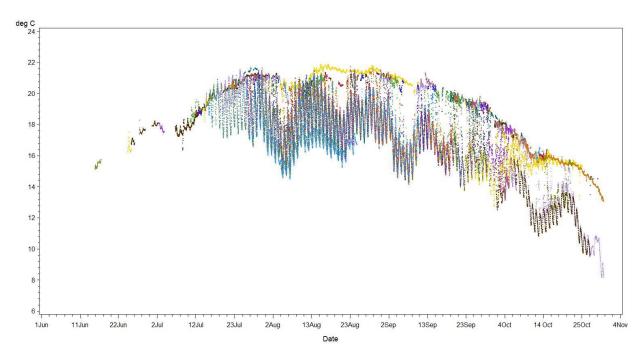


Figure 16. Thermal histories (30-min) of RDST-tagged steelhead (n = 117) in The Dalles reservoir in 2002.

3.5.6 John Day reservoir reach: Chinook salmon

A total of 67 summer Chinook salmon had RDST data (Table 9). They all passed McNary Dam with no evidence of CWR use (Figures 17 and 18). Thirty-nine fall Chinook salmon had RDST data and 30 (77%) of these passed John Day Dam. A large majority (92%, n = 36) had no evidence of CWR use. One each had incidental use < 4 h, use of a single CWR (Umatilla River), and use of an unknown CWR site (possibly Irrigon Hatchery outfall). A single fall Chinook salmon was last detected in the Umatilla River (Table 3).

Table 9. Numbers of summer and fall Chinook salmon with RDST temperature data collected in the John Day pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the McNary Dam WQM site were used as the ambient temperature reference because John Day WOM data were not available through the fall.

	uala were not a	Passed	<u> </u>	Cold water re	efuge (CWR)	use		
Year	Run	McNary	п	None	Incidental	Unk CWR	1 CWR	>1 CWR
2000	Summer	Yes	37	37 (100%)				
		No	-					
2002	Summer	Yes	30	30 (100%				
		No	-					
Both	Summer-W	Yes+No	13	13 (100%)				
Dom								
	Summer-H	Yes+No	54	54 (100%)				
Both	$T \ge 20^{\circ}C$	Yes+No	9	9 (100%)				
Both	$T < 20^{\circ}C$	Yes+No	58	58 (100%				
2000	Fall	Yes	15	15 (100%)				
		No	6	4 (67%)		2 (33%)		
		n/a ¹	5					
2002	Fall	Yes	15	14 (93%)	1 (7%)			
		No	3	3 (100%)				
Detk		VactNa	16	15 (040)	1(60/)			
Both	Fall-W	Yes+No	16	15 (94%)	1 (6%)	2 (00()		
	Fall-H	Yes+No	23	21 (91%)		2 (9%)		
Both	$T \ge 20^{\circ}C$	Yes+No	29	28 (97%)		1 (3%)		
Both	$T < 20^{\circ}C$	Yes+No	10	8 (80%)	1 (10%)	1 (10%)		

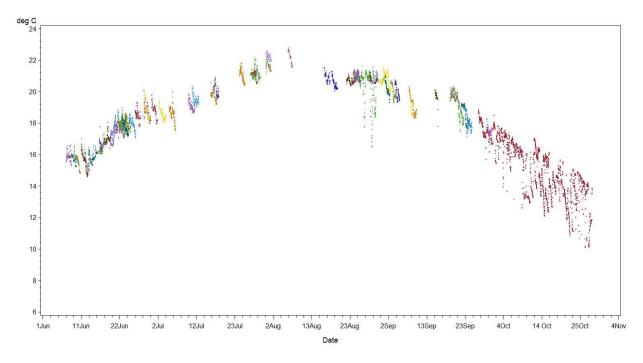


Figure 17. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 37) and fall Chinook salmon (n = 22) in John Day reservoir in 2000.

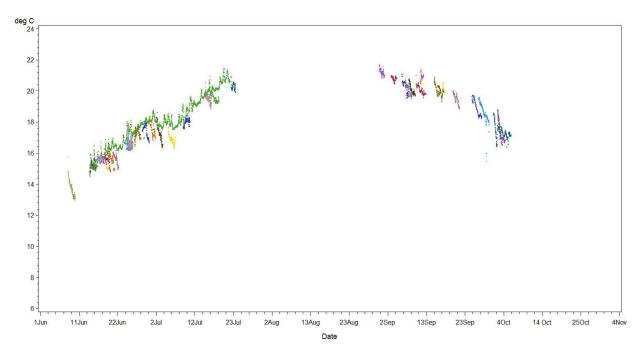


Figure 18. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 37) and fall Chinook salmon (n = 16) in John Day reservoir in 2002.

3.5.7 John Day reservoir reach: Steelhead

A total of 136 steelhead had RDST data in the reach and 129 (95%) of these passed McNary Dam (Table 10). A large majority (90%, n = 123) had no evidence of CWR use (Figures 19 and 20). Eight steelhead (6% of 136) had incidental CWR use and 5 (4%) used unknown CWR sites. Potential CWR locations included the John Day and Umatilla rivers, Rock Creek, the Irrigon Hatchery outfall, or other unidentified sites. A single steelhead was last detected in the John Day River (Table 3 and Appendix A). Patterns were similar among wild and hatchery steelhead and for the encountered temperature groups.

Table 10. Numbers of steelhead with RDST temperature data collected in the John Day pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the McNary Dam WQM site were used as the ambient temperature reference because John Day WQM data were not available through the fall.

		Passed		Cold water re	efuge (CWR)	use		
Year	Run	McNary	n	None	Incidental	Unk CWR	1 CWR	>1 CWR
2000	Steelhead	Yes	45	42 (93%)	2 (4%)	1 (2%)		
		No	5	5 (100%)				
		n/a^1	21					
2002	Steelhead	Yes	84	75 (89%)	5 (6%)	4 (5%)		
		No	2	1 (50%)	1 (50%)			
		n/a^1	53					
Both	Steelhead-W	Yes+No	75	65 (87%)	6 (8%)	4 (5%)		
	Steelhead-H	Yes+No	62	59 (95%)	2 (3%)	1 (2%)		
Both	$T \ge 20^{\circ}C$	Yes+No	26	25 (96%)	1 (4%)			
Both	$T < 20^{\circ}C$	Yes+No	111	99 (89%)	7 (6%)	5 (5%)		

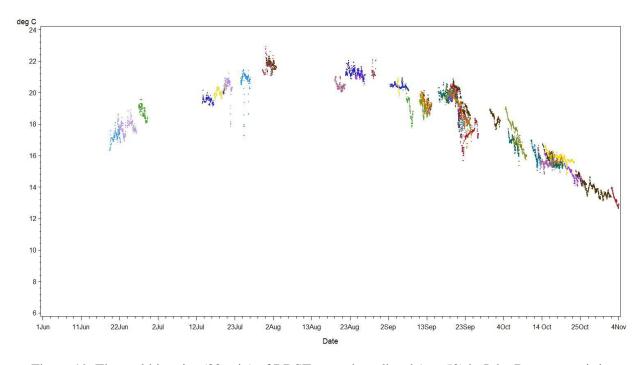


Figure 19. Thermal histories (30-min) of RDST-tagged steelhead (n = 52) in John Day reservoir in 2000.

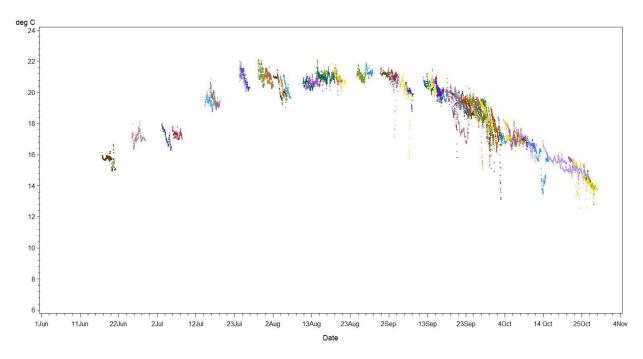


Figure 20. Thermal histories (30-min) of RDST-tagged steelhead (n = 86) in John Day reservoir in 2002.

3.5.8 McNary reservoir reach: Chinook salmon

A total of 66 summer Chinook salmon had RDST data (Table 11). They all passed either Ice Harbor Dam or into the Hanford Reach. A large majority (94%, n = 62) had no evidence of CWR use (Figures 21 and 22). Four fish (6%) used and unknown CWR site, potentially the Walla Walla River, though no fish were detected on the Walla Walla antenna (perhaps due to the site configuration). It is also possible that some fish used cooler water near the Columbia River–Snake River confluence; no antenna monitored the transition from McNary pool into the free-flowing Hanford Reach. In contrast to behavior in downstream reaches, summer Chinook that encountered reservoir temperatures ≥ 20 °C were slightly less likely to use a CWR site, but only 8 fish were in the warm category.

Thirty fall Chinook salmon had RDST data and all passed through the reach (Table 11). All but one salmon (97%, n = 29) had no evidence of CWR use (Figures 21 and 22). The single fish with CWR use was at an unknown location and did not encounter temperatures ≥ 20 °C.

	Passed Cold water refuge (CWR) use							
Year	Run	MCN pool	n	None	Incidental	Unk CWR	1 CWR	>1 CWR
2000	Summer	Yes	37	34 (92%)		3 (8%)		
		No	-					
2002	Summer	Yes	29	28 (97%)		1 (3%)		
		No	-					
		n/a ¹	1					
Both	Summer-W	Yes+No	11	11 (100%)				
	Summer-H	Yes+No	55	51 (93%)		4 (7%)		
Both	$T \ge 20^{\circ}C$	Yes+No	8	8 (100%)				
Both	$T < 20^{\circ}C$	Yes+No	58	54 (93%)		4 (7%)		
• • • • •	5 11					4 (7 8 ()		
2000	Fall	Yes	15	14 (93%)		1 (7%)		
		No	-					
2002	E-11	n/a^1	2	15(1000/)				
2002	Fall	Yes No	15	15 (100%)				
		NO	-					
Both	Fall-W	Yes+No	13	13 (100%)				
	Fall-H	Yes+No	18	17 (94%)		1 (6%)		
Both	$T \ge 20^{\circ}C$	Yes+No	6	6 (100%)				
Both	$T \ge 20^{\circ} C$ $T < 20^{\circ} C$	Yes+No	25	24 (96%)		1 (4%)		

Table 11. Numbers of summer and fall Chinook salmon with RDST temperature data collected in the McNary pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the McNary Dam WQM site were used as the ambient temperature reference.

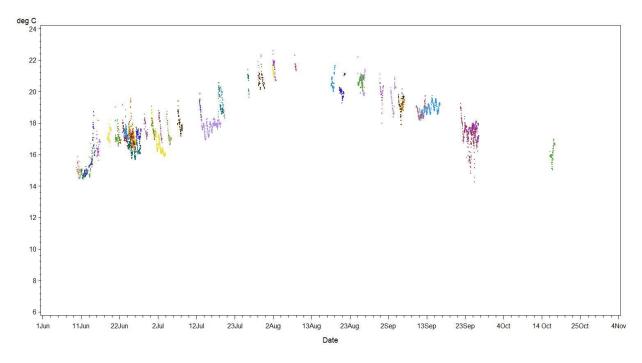


Figure 21. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 37) and fall Chinook salmon (n = 16) in McNary reservoir in 2000.

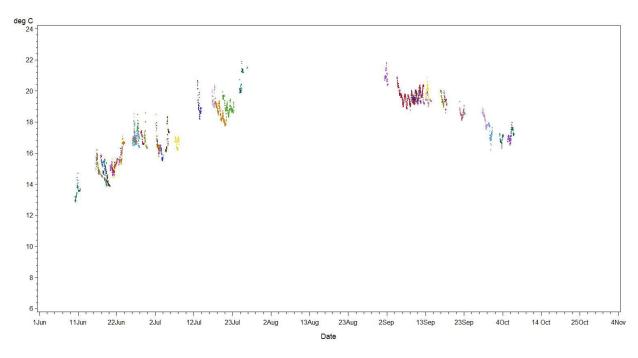


Figure 22. Thermal histories (30-min) of RDST-tagged summer Chinook salmon (n = 29) and fall Chinook salmon (n = 15) in McNary reservoir in 2002.

3.5.9 McNary reservoir reach: Steelhead

A total of 117 steelhead had RDST data in the reach and 112 (96%) of these passed through the reach (Table 12). Several of the non-passers were reported harvested in the reservoir. A majority (80%, n = 94) of the steelhead had no evidence of CWR use (Figures 23 and 24). Patterns were similar for wild and hatchery fish. In contrast to in downstream reaches, steelhead that encountered reservoir temperatures ≥ 20 °C were somewhat less likely to use CWR sites. Thirteen steelhead (11% of 117) had incidental CWR use and 10 (9%) used unknown CWR sites. Potential CWR sites included the Walla Walla River, the lower end of the Hanford Reach, or other unidentified sites. A single steelhead was last detected in the Walla Walla River (Table 4).

Table 12. Numbers of steelhead with RDST temperature data collected in the McNary pool reach, classified by their use / non-use of cold water refuge (CWR) sites, by hatchery (H) or presumed wild (W) origin, and whether they encountered water temperature $\geq 20^{\circ}$ C. Temperature data from the McNary Dam WQM site were used as the ambient temperature reference.

		Passed		CWR use				
Year	Run	MCN pool	п	None	Incidental	Unk CWR	1 CWR	>1 CWR
2000	Steelhead	Yes	39	34 (87%)	3 (8%)	2 (5%)		
		No	4	4 (100%)				
		n/a^1	20					
2002	Steelhead	Yes	73	55 (75%)	10 (14%)	8 (11%)		
		No	1	1 (100%)				
		n/a ¹	61					
Both	Steelhead-W	Yes+No	64	49 (77%)	7 (11%)	8 (13%)		
	Steelhead-H	Yes+No	53	45 (85%)	6 (11%)	2 (4%)		
Both	$T \ge 20^{\circ}C$	Yes+No	48	43 (90%)	3 (6%)	2 (4%)		
Both	$T < 20^{\circ}C$	Yes+No	69	51 (74%)	10 (14%)	8 (12%)		

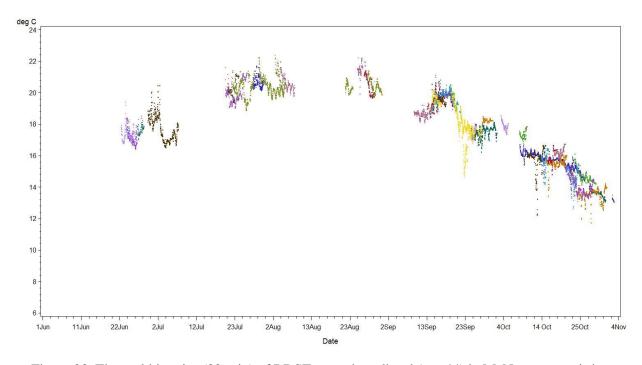


Figure 23. Thermal histories (30-min) of RDST-tagged steelhead (n = 44) in McNary reservoir in 2000.

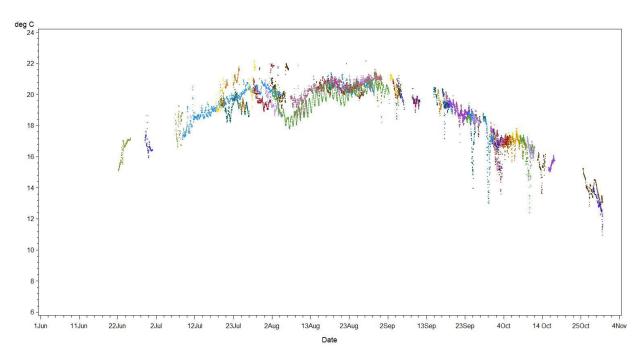


Figure 24. Thermal histories (30-min) of RDST-tagged steelhead (n = 74) in McNary reservoir in 2002.

3.6 Residence times in each reach

Reach residence times were calculated using the number of RDST records (30-min) assigned to each reach based on the radiotelemetry data (Figure 25). Time that fish spent in reservoir reaches included their time in CWR sites for this section. Time that fish spent at dams included time spent in tailraces and in fishways. The RDST-based estimates should not be considered equivalent to passage times based strictly on radiotelemetry data at the start and end of a reach for several reasons. First, some fish were in a reach for a short amount of time before they were harvested (i.e., they did not pass entirely through the reach). Second, it was not always possible to infer when fish moved downstream from tailraces back into reservoirs. Third, it was also difficult to determine when fish exited the McNary reservoir reach into the Hanford Reach, as there was no antenna monitoring the transition between these habitats. Fourth, in some cases the RDST storage capacity was exceeded while a fish was traversing a reach, and hence residence time was underestimated. Consequently, the summaries in Figure 25 should be a good representation of the duration of the temperature data collected in each reach, but there was some estimation error of total residence time for some fish in each run. For a thorough summary of salmon and steelhead passage times through the study reaches, with much larger sample sizes, see Keefer et al. (2004, 2009) and Goniea et al. (2006).

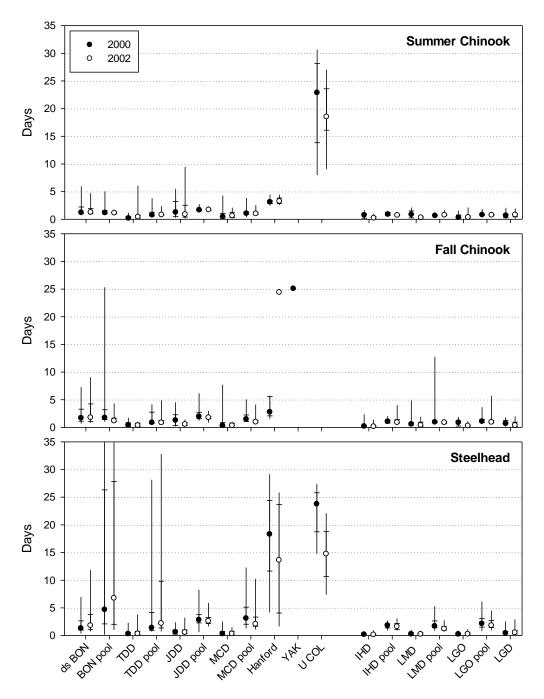


Figure 25. Distributions of time RDST-tagged fish collected temperature data in each reach in 2000 (•) and 2002 (•). Whisker plots show 5th, 25th, 50th, 75th, and 95th percentiles. BON = Bonneville, TDD = The Dalles, JDD = John Day, MCD = McNary, YAK = Yakima River, U COL = upper Columbia River, IHD = Ice Harbor, LMD = Lower Monumental, LGO = Little Goose, LGD = Lower Granite. Note: a small number of steelhead times in the Bonneville pool are not shown.

Most summer Chinook salmon spent < 4 d in each lower Columbia reach (Figure 25), reflecting generally rapid upstream migration with little CWR use. Median migration times for fall Chinook salmon were also < 4 d in each lower river reach, although the passage time distribution was right-skewed in the Bonneville pool reach relative to summer Chinook salmon (Figure 25). Steelhead passage times were far more variable than those of Chinook salmon, with some fish taking weeks to pass through a reach; however, median steelhead residence times based on RDST temperature records were ≤ 10 d in all lower river reaches.

3.7 Migration rates in reservoir reaches

To better standardize comparisons across reaches and to inform the HexSim model timesteps, we calculated passage rates (km/h) for RDST-tagged fish through the four reservoir reaches. Start times were the last radiotelemetry detection at a downstream dam (i.e., fishway exit antenna) to the first detection at an upstream dam (i.e., fishway approach antenna). Fish that were harvested in the reach or that were last detected in tributaries in the reach were excluded from rate calculations. For a thorough summary of salmon and steelhead passage rates through the study reaches, with much larger sample sizes, see Keefer et al. (2004, 2009) and Goniea et al. (2006).

3.7.1 Bonneville reservoir reach

Summer Chinook salmon passed through the Bonneville pool at rates ranging from 1.13-3.49 km/h (*mean* = 2.55, *median* = 2.67) (Figure 26). Fall Chinook salmon passed through the Bonneville pool at rates ranging from 0.03 km/h (a very slow fish that took 33 d to pass the reach) to 3.28 km/h (*mean* = 2.31, *median* = 2.44) (Figure 27). Steelhead passage rates were left-skewed, reflecting extensive CWR use in the Bonneville pool reach. Steelhead rates ranged from 0.10-2.44 km/h (*mean* = 0.79, *median* = 0.58) (Figure 28).

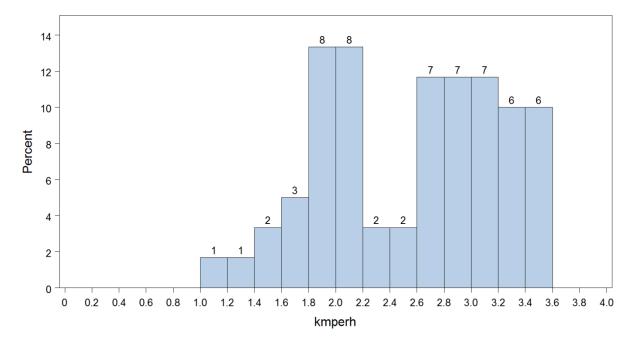


Figure 26. Histogram of summer Chinook salmon passage rates (km/h) from fishway exit at Bonneville Dam to first fishway approach at The Dalles Dam, 2000 and 2002. Number above bar = n.

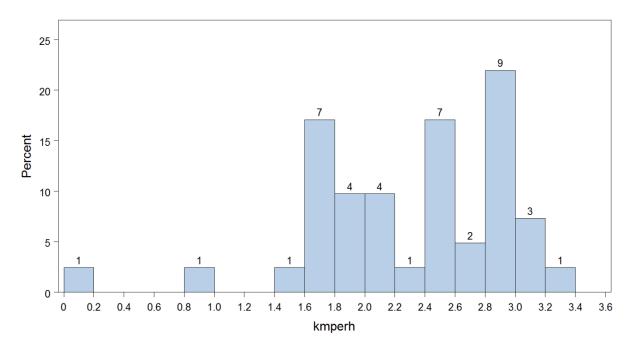


Figure 27. Histogram of fall Chinook salmon passage rates (km/h) from fishway exit at Bonneville Dam to first fishway approach at The Dalles Dam, 2000 and 2002. Number above bar = n.

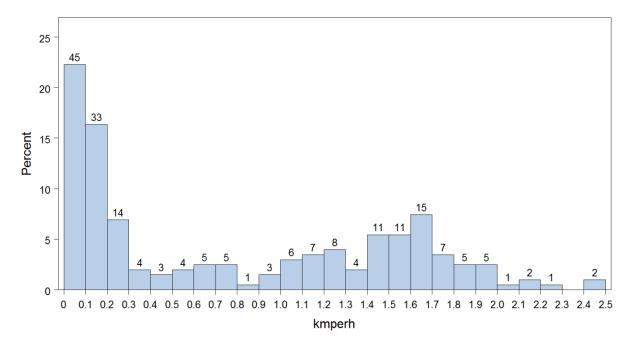


Figure 28. Histogram of steelhead passage rates (km/h) from fishway exit at Bonneville Dam to first fishway approach at The Dalles Dam, 2000 and 2002. Number above bar = n.

3.7.2 The Dalles reservoir reach

Summer Chinook salmon passed through the The Dalles pool at rates ranging from 0.24-3.66 km/h (*mean* = 2.34, *median* = 2.29) (Figure 29). Fall Chinook salmon rates ranged from 0.11-3.88 km/h (*mean* = 1.96, *median* = 2.05) (Figure 30). Steelhead passage rates were much more variable, reflecting extensive use of the Deschutes River CWR. Rates ranged from 0.01-2.63 km/h (*mean* = 0.98, *median* = 1.04) (Figure 31).

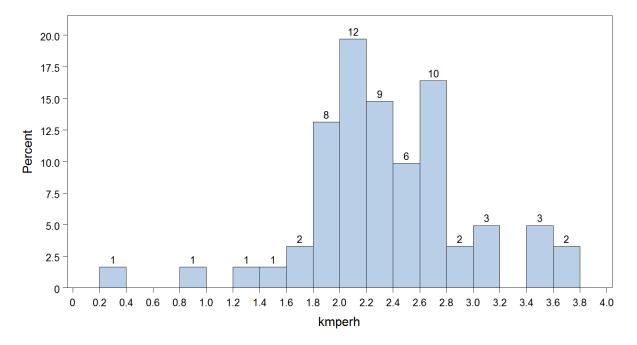


Figure 29. Histogram of summer Chinook salmon passage rates (km/h) from fishway exit at The Dalles Dam to first fishway approach at John Day Dam, 2000 and 2002. Number above bar = n.

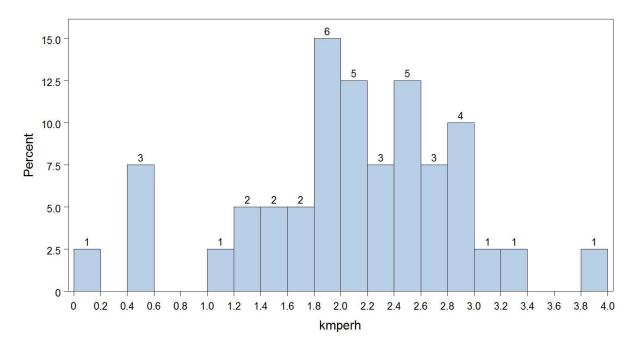


Figure 30. Histogram of fall Chinook salmon passage rates (km/h) from fishway exit at The Dalles Dam to first fishway approach at John Day Dam, 2000 and 2002. Number above bar = n.

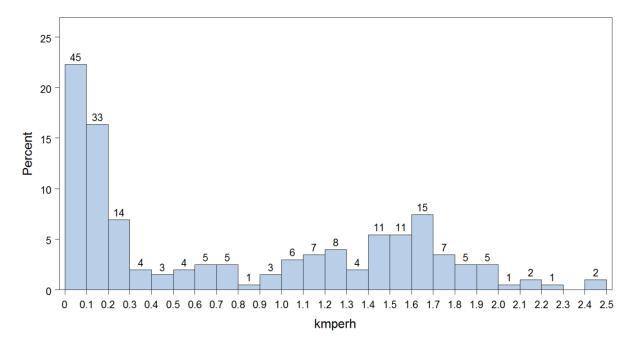


Figure 31. Histogram of steelhead passage rates (km/h) from fishway exit at The Dalles Dam to first fishway approach at John Day Dam, 2000 and 2002. Number above bar = n.

3.7.3 John Day reservoir reach

Summer Chinook salmon passed through the John Day pool at rates ranging from 0.10-4.27 km/h (*mean* = 2.95, *median* = 3.03) (Figure 32). Fall Chinook salmon rates ranged from 1.39–3.67 km/h (*mean* = 2.76, *median* = 2.78) (Figure 33). Steelhead passage rates ranged from 0.09–3.19 km/h (*mean* = 1.86, *median* = 1.90) (Figure 34).

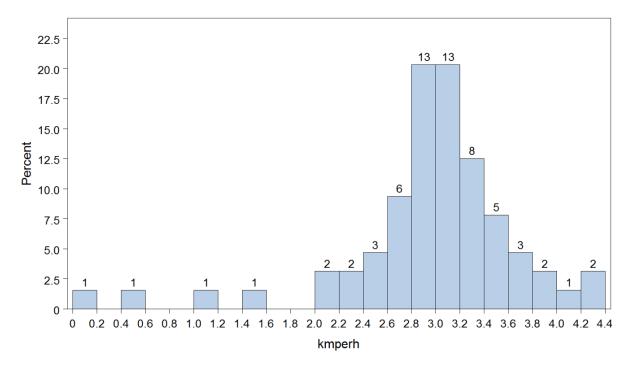


Figure 32. Histogram of summer Chinook salmon passage rates (km/h) from fishway exit at John Day Dam to first fishway approach at McNary Dam, 2000 and 2002. Number above bar = n.

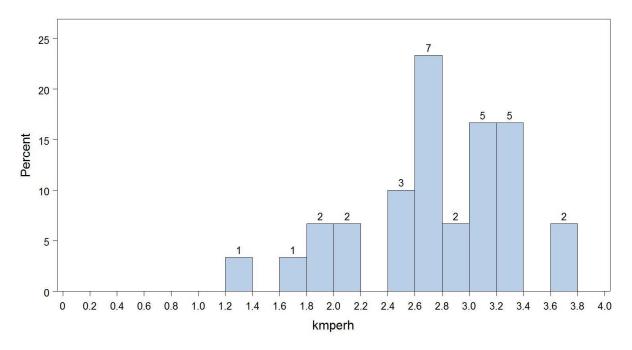


Figure 33. Histogram of fall Chinook salmon passage rates (km/h) from fishway exit at John Day Dam to first fishway approach at McNary Dam, 2000 and 2002. Number above bar = n.

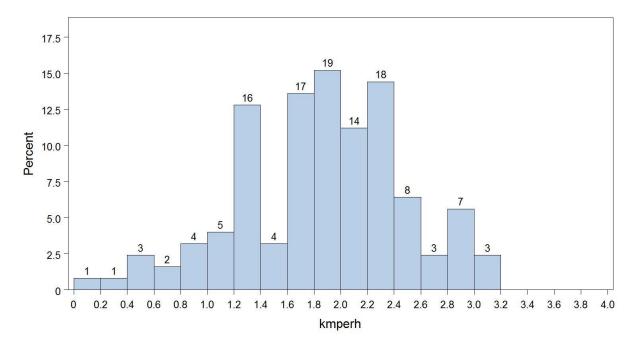


Figure 34. Histogram of steelhead passage rates (km/h) from fishway exit at John Day Dam to first fishway approach at McNary Dam, 2000 and 2002. Number above bar = n.

3.7.4 McNary reservoir reach

Summer Chinook salmon passed through the John Day pool at rates ranging from 1.36-3.73 km/h (*mean* = 2.80, *median* = 2.94) (Figure 35). Fall Chinook salmon rates ranged from 0.61– 3.57 km/h (*mean* = 2.42, *median* = 2.68) (Figure 36). Steelhead passage rates ranged from 0.02– 2.73 km/h (*mean* = 1.21, *median* = 1.31) (Figure 34). (Note: calculations for the McNary pool reach do not include rates for fish that continued migrating up the Columbia River as there was no antenna to monitor reservoir exit).

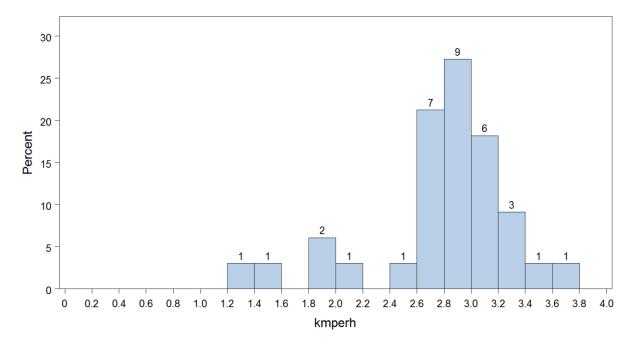


Figure 35. Histogram of summer Chinook salmon passage rates (km/h) from fishway exit at McNary Dam to first fishway approach at Ice Harbor Dam, 2000 and 2002. Number above bar = n.

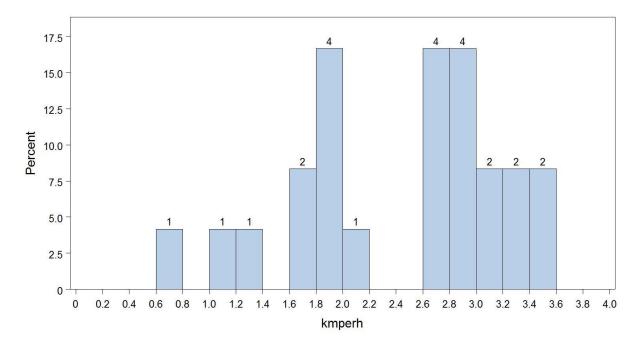


Figure 36. Histogram of fall Chinook salmon passage rates (km/h) from fishway exit at McNary Dam to first fishway approach at Ice Harbor Dam, 2000 and 2002. Number above bar = n.

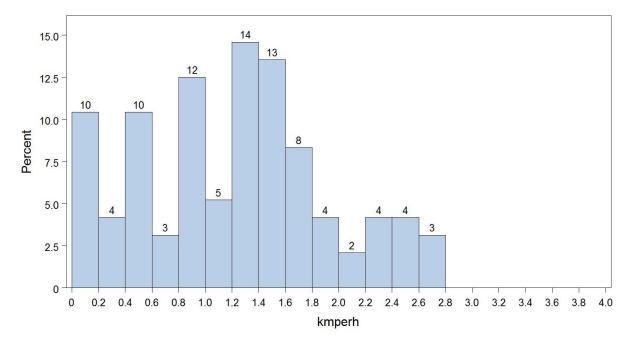


Figure 37. Histogram of steelhead passage rates (km/h) from fishway exit at McNary Dam to first fishway approach at Ice Harbor Dam, 2000 and 2002. Number above bar = n.

3.8 Residence times in cold water refuge (CWR) sites

Time RDST-tagged fish spent in cold water refuge (CWR) sites was calculated using the criteria of ≤ 2 °C below ambient reservoir temperature. We did not differentiate among CWR sites for this summary. Fish that were harvested in the reach or that were last detected in tributaries in the reach were excluded from calculations (i.e., only includes fish that passed the reach).

3.8.1 Bonneville reservoir reach

Most summer Chinook salmon that passed The Dalles Dam had no CWR records in the Bonneville reservoir and none had > 1 d of use (Figure 38). The longest CWR use by a summer Chinook salmon was 0.5 d. Twelve fall Chinook salmon had CWR use > 4 h (*mean* = 2.4 d, *median* = 1.2 d) (Figure 39). The distribution of steelhead CWR use times was much more variable (Figure 40). In total, 141 steelhead had CWR use > 4 h (*mean* =19.5 d, *median* = 20.8 d). Some steelhead also reached RDST storage limits while in Bonneville CWR sites, indicating that mean, median and maximum values were likely underestimated.

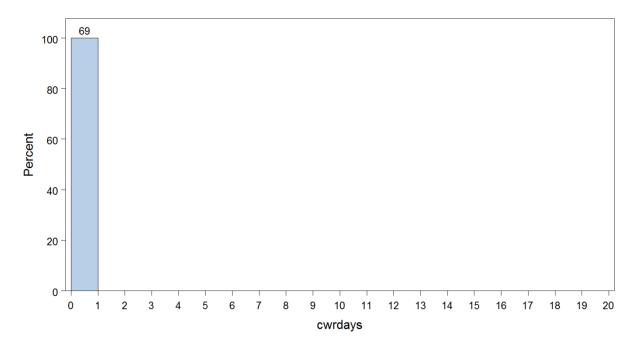


Figure 38. Histogram of summer Chinook salmon residence times in cold water refuge sites (cwrdays) in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

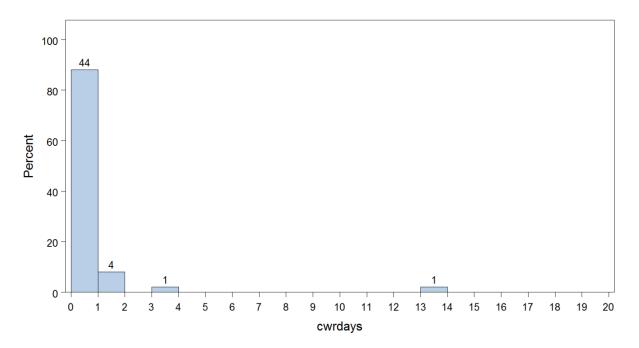


Figure 39. Histogram of fall Chinook salmon residence times in cold water refuge sites (cwrdays) in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

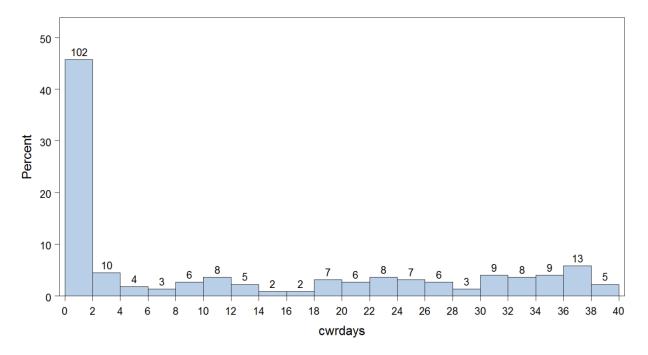


Figure 40. Histogram of steelhead residence times in cold water refuge sites (cwrdays) in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

3.8.2 The Dalles reservoir reach

Most summer Chinook salmon that passed John Day Dam had no CWR records in The Dalles reservoir and only one had > 1 d of use (6.9 d, Figure 41). Ten fall Chinook salmon had CWR use > 4 h (*mean* = 0.9 d, *median* = 0.6 d) (Figure 42). The distribution of steelhead CWR use times was much more variable (Figure 43). In total, 93 steelhead had CWR use > 4 h (*mean* = 6.3 d, *median* = 2.1 d). Some steelhead also reached RDST storage limits while in the reach, indicating that mean, median and maximum values were likely underestimated.

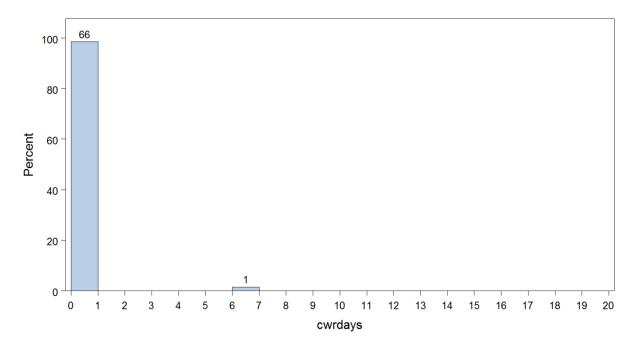


Figure 41. Histogram of summer Chinook salmon residence times in cold water refuge sites (cwrdays) in The Dalles pool reach, 2000 and 2002. Number above bar = n.

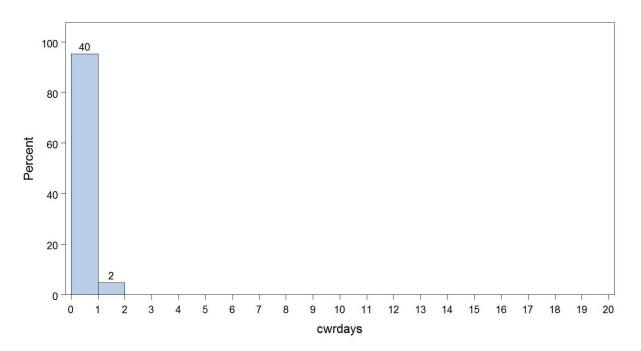


Figure 42. Histogram of fall Chinook salmon residence times in cold water refuge sites (cwrdays) in The Dalles pool reach, 2000 and 2002. Number above bar = n.

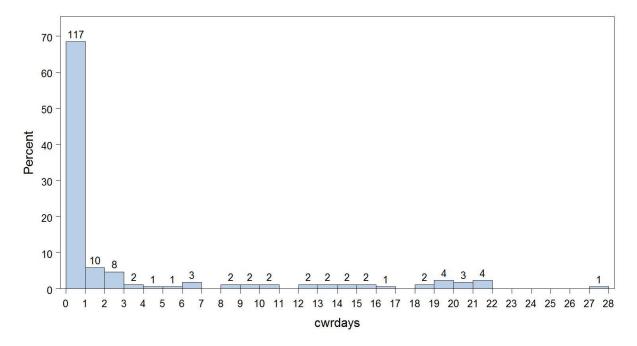


Figure 43. Histogram of steelhead residence times in cold water refuge sites (cwrdays) in The Dalles pool reach, 2000 and 2002. Number above bar = n.

3.8.3 John Day reservoir reach

Of the fish that passed McNary Dam, no summer or fall Chinook salmon used a CWR site in the John Day reservoir reach for > 4 h (Figure 44-45). Five steelhead had CWR use > 4 h, but none had use > 1 d (Figure 46).

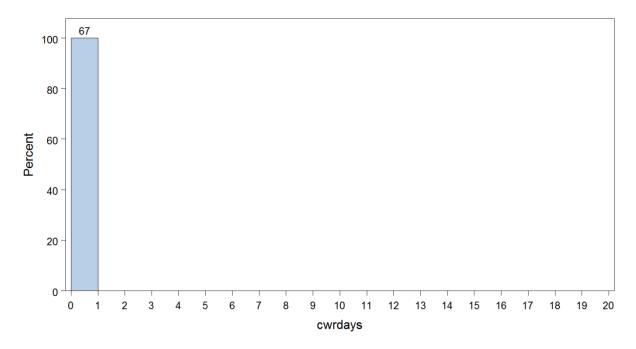


Figure 44. Histogram of summer Chinook salmon residence times in cold water refuge sites (cwrdays) in the John Day pool reach, 2000 and 2002. Number above bar = n.

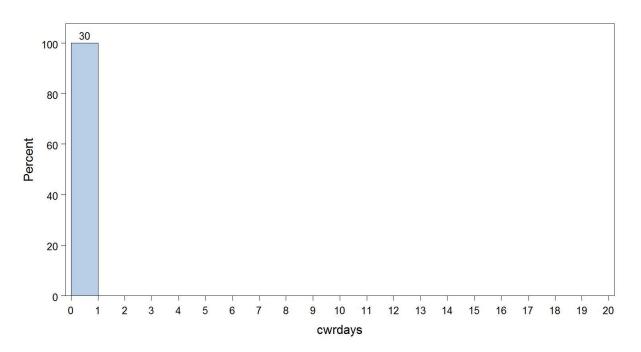


Figure 45. Histogram of fall Chinook salmon residence times in cold water refuge sites (cwrdays) in the John Day pool reach, 2000 and 2002. Number above bar = n.

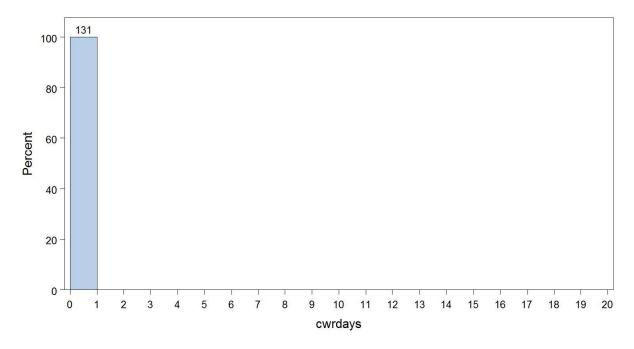


Figure 46. Histogram of steelhead residence times in cold water refuge sites (cwrdays) in the John Day pool reach, 2000 and 2002. Number above bar = n.

3.8.4 McNary reservoir reach

Of the fish that passed Ice Harbor Dam, three summer Chinook salmon used a CWR site in McNary reservoir for > 4 h (*mean* = 0.8 d, *median* = 0.9 d) (Figure 47). A single fall Chinook salmon used a CWR site > 4 h (0.4 d, Figure 48). Ten steelhead had CWR use > 4 h, but none had use > 1 d (Figure 49).

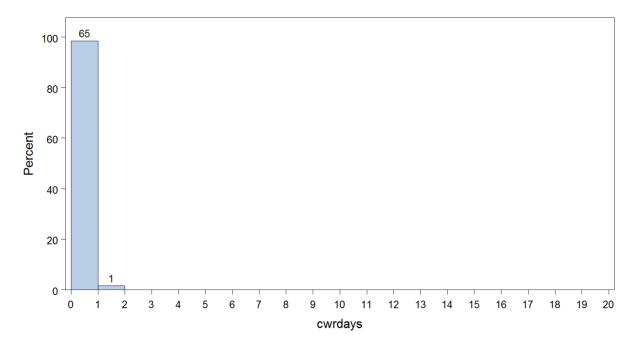


Figure 47. Histogram of summer Chinook salmon residence times in cold water refuge sites (cwrdays) in the McNary pool reach, 2000 and 2002. Number above bar = n.

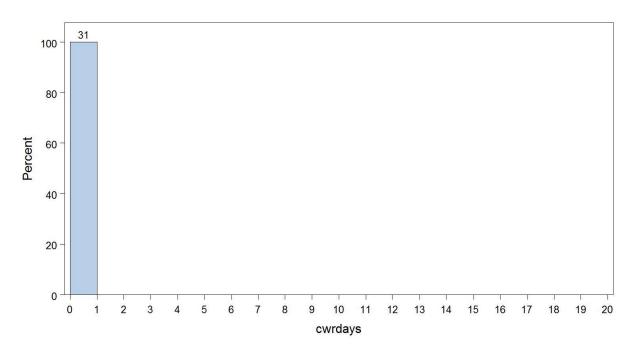


Figure 48. Histogram of fall Chinook salmon residence times in cold water refuge sites (cwrdays) in the McNary pool reach, 2000 and 2002. Number above bar = n.

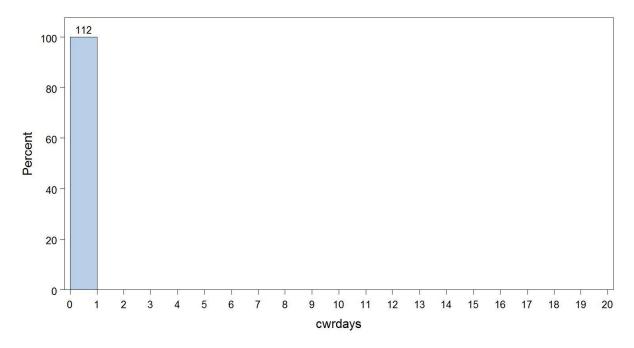


Figure 49. Histogram of steelhead residence times in cold water refuge sites (cwrdays) in the McNary pool reach, 2000 and 2002. Number above bar = n.

3.9 Mean temperature exposure per reach

Mean body temperatures for the RDST-tagged fish were calculated for each reservoir reach. These estimates included all times and locations in each reach, including time spent in CWR sites. Fish that were harvested in the reach or that were last detected in tributaries in the reach were excluded from calculations. Distributions of mean temperature exposure may be a useful target metric for HexSim model outputs.

3.9.1 Bonneville reservoir reach

Mean temperatures of summer Chinook salmon that passed The Dalles Dam ranged from 14.5 °C to 21.6 °C (*mean of means* = 17.0 °C, *median of means* = 16.6 °C) (Figure 50). Means for fall Chinook salmon ranged from 12.1 °C to 21.3 °C (*mean of means* = 19.4 °C, *median of means* = 19.7 °C) (Figure 50). The small number of fall Chinook salmon with means < 18 °C included two very late-run migrants and two with extended CWR use. Means for steelhead were lower and more variable than for both Chinook salmon runs (Figure 51), reflecting much higher incidence of CWR use. The range was from 9.8 °C to 21.1 °C (*mean of means* = 16.2 °C, *median of means* = 15.8 °C).

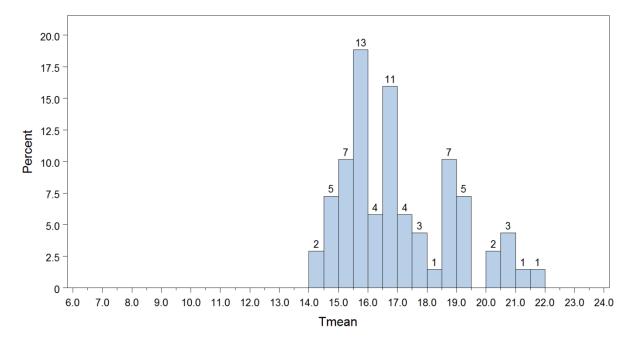


Figure 50. Mean body temperature of RDST-tagged summer Chinook salmon in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

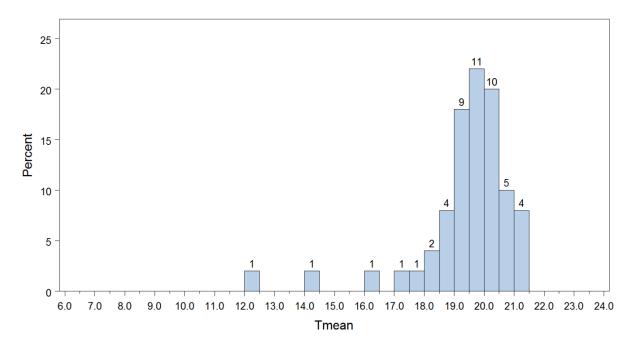


Figure 51. Mean body temperature of RDST-tagged fall Chinook salmon in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

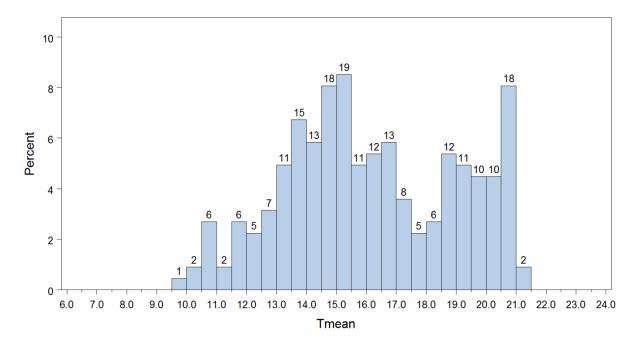


Figure 52. Mean body temperature of RDST-tagged steelhead in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

3.9.2 The Dalles reservoir reach

Mean temperatures of summer Chinook salmon that passed John Day Dam ranged from 14.2 °C to 22.0 °C (*mean of means* = 17.2 °C, *median of means* = 16.5 °C) (Figure 53). Means for fall Chinook salmon ranged from 12.8 °C to 21.1 °C (*mean of means* = 19.5 °C, *median of means* = 20.0 °C) (Figure 54). The two fall Chinook salmon with means < 17 °C were very late-run migrants. Means for steelhead were lower and more variable than for both Chinook salmon runs (Figure 55), reflecting much higher use of the Deschutes River CWR. The range was from 11.5 °C to 21.4 °C (*mean of means* = 18.2 °C, *median of means* = 18.3 °C).

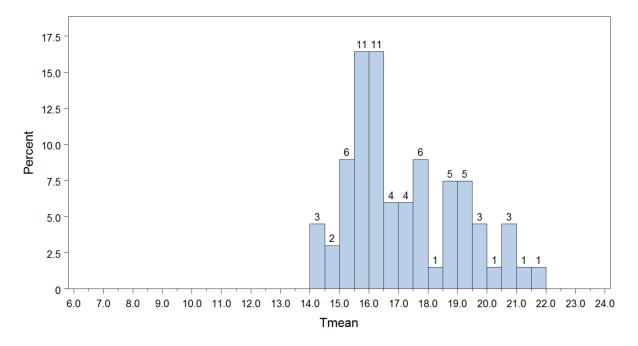


Figure 53. Mean body temperature of RDST-tagged summer Chinook salmon in The Dalles pool reach, 2000 and 2002. Number above bar = n.

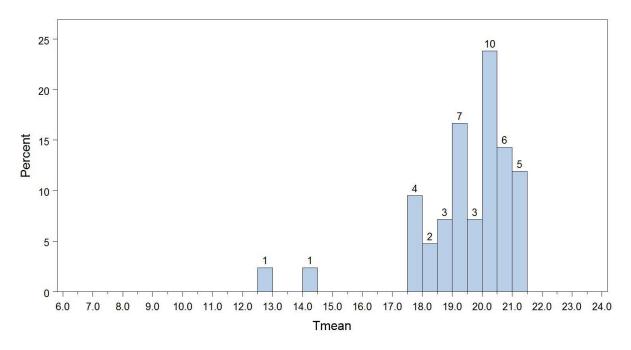


Figure 54. Mean body temperature of RDST-tagged fall Chinook salmon in The Dalles pool reach, 2000 and 2002. Number above bar = n.

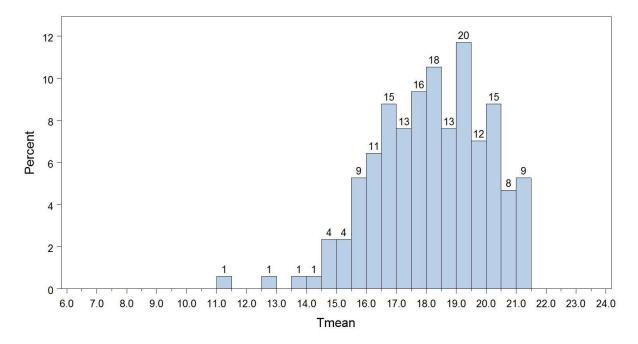


Figure 55. Mean body temperature of RDST-tagged steelhead in The Dalles pool reach, 2000 and 2002. Number above bar = n.

3.9.3 John Day reservoir reach

Mean temperatures of summer Chinook salmon that passed McNary Dam ranged from 13.8 °C to 22.2 °C (*mean of means* = 17.6 °C, *median of means* = 17.4 °C) (Figure 56). Means for fall Chinook salmon ranged from 16.0 °C to 21.7 °C (*mean of means* = 19.6 °C, *median of means* = 20.0 °C) (Figure 57). Means for steelhead were somewhat more variable than for both Chinook salmon runs (Figure 58); the lowest means were for migrants in late October and early November. The range for steelhead was from 11.0 °C to 21.8 °C (*mean of means* = 18.7 °C, *median of means* = 19.1 °C).

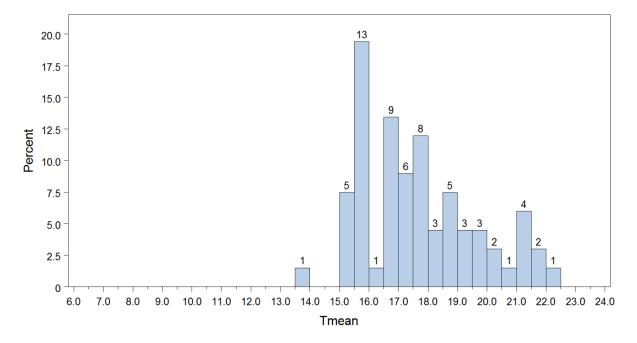


Figure 56. Mean body temperature of RDST-tagged summer Chinook salmon in the John Day pool reach, 2000 and 2002. Number above bar = n.

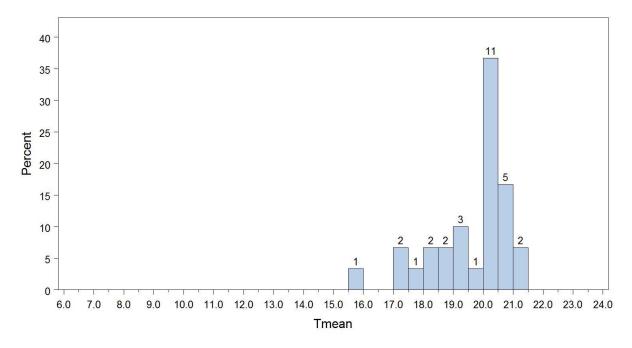


Figure 57. Mean body temperature of RDST-tagged fall Chinook salmon in the John Day pool reach, 2000 and 2002. Number above bar = n.

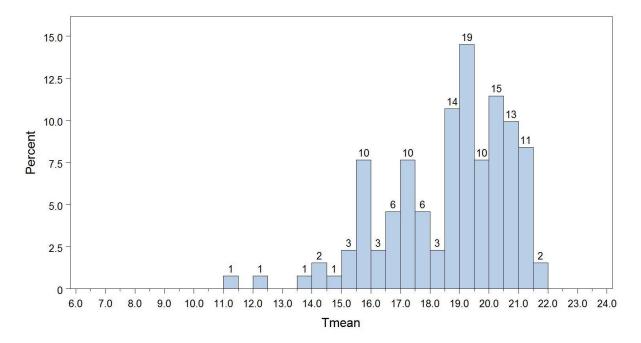


Figure 58. Mean body temperature of RDST-tagged steelhead in the John Day pool reach, 2000 and 2002. Number above bar = n.

3.9.4 McNary reservoir reach

Mean temperatures of summer Chinook salmon that passed through the McNary reach ranged from 13.5 °C to 21.6 °C (*mean of means* = 17.3 °C, *median of means* = 17.0 °C) (Figure 59). Means for fall Chinook salmon ranged from 11.7 °C to 20.9 °C (*mean of means* = 18.7 °C, *median of means* = 19.2 °C) (Figure 60). The lowest mean was for a mid-November migrant. The range for steelhead was bimodally distributed (Figure 61) from 11.5 °C to 21.5 °C (*mean of means* = 18.2 °C, *median of means* = 18.6 °C). The cooler mode included a higher proportion of late-run migrants than the warmer mode.

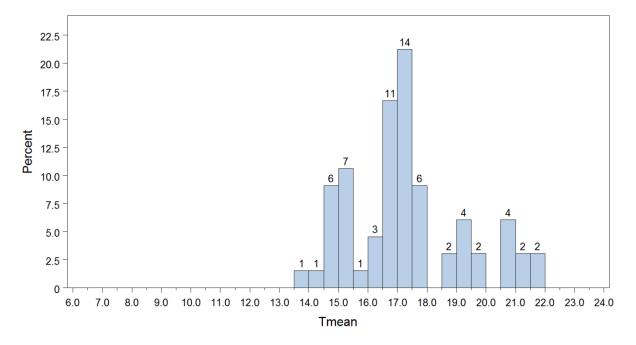


Figure 59. Mean body temperature of RDST-tagged summer Chinook salmon in the McNary pool reach, 2000 and 2002. Number above bar = n.

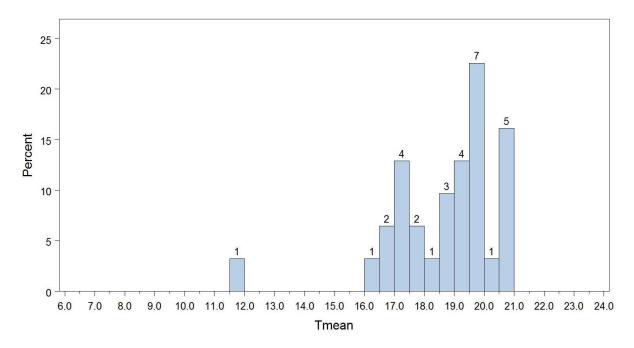


Figure 60. Mean body temperature of RDST-tagged fall Chinook salmon in the McNary pool reach, 2000 and 2002. Number above bar = n.

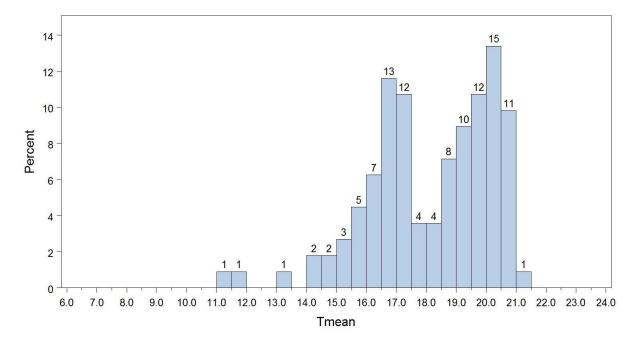


Figure 61. Mean body temperature of RDST-tagged steelhead in the McNary pool reach, 2000 and 2002. Number above bar = n.

3.10 Maximum temperature exposure per reach

Maximum body temperatures for the RDST-tagged fish were calculated for each reservoir reach. These estimates included all times and locations in each reach. Fish that were harvested in the reach or that were last detected in tributaries in the reach were excluded from calculations. Distributions of maximum temperature exposure may be a useful target metric for HexSim model outputs.

3.10.1 Bonneville reservoir reach

Maximum temperatures of summer Chinook salmon that passed The Dalles Dam ranged from 14.8 °C to 22.2 °C (*mean of maximums* = 17.5 °C, *median of maximums* = 17.1 °C) (Figure 62). Means for fall Chinook salmon ranged from 14.7 °C to 21.6 °C (*mean of maximums* = 20.2 °C, *median of maximums* = 20.5 °C) (Figure 63). Maximums for steelhead ranged from 15.7 °C to 22.4 °C (*mean of maximums* = 20.2 °C, *median of maximums* = 20.7 °C). Low maximum values for steelhead tended to be early or late in the runs.

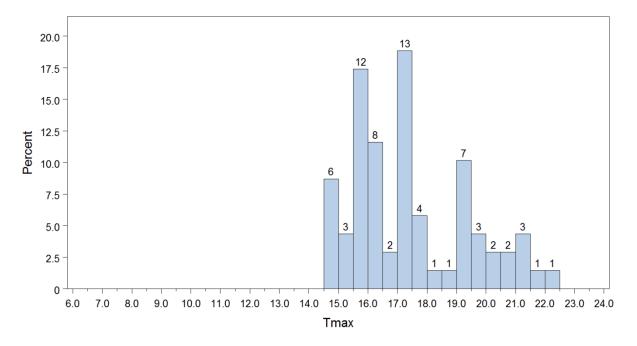


Figure 62. Maximum body temperature of RDST-tagged summer Chinook salmon in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

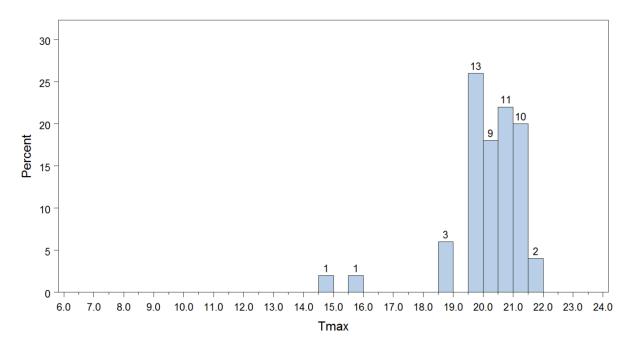


Figure 63. Maximum body temperature of RDST-tagged fall Chinook salmon in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

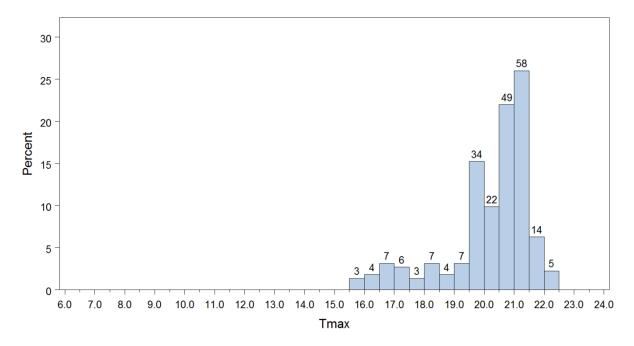


Figure 64. Maximum body temperature of RDST-tagged steelhead in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

3. 10.2 The Dalles reservoir reach

Maximum temperatures of summer Chinook salmon that passed John Day Dam ranged from 14.6 °C to 22.5 °C (*mean of maximums* = 17.6 °C, *median of maximums* = 16.9 °C) (Figure 65). Means for fall Chinook salmon ranged from 13.0 °C to 21.5 °C (*mean of maximums* = 20.1 °C, *median of maximums* = 20.5 °C) (Figure 66). Maximums for steelhead ranged from 14.6 °C to 22.3 °C (*mean of maximums* = 19.7 °C, *median of maximums* = 20.6 °C) (Figure 67). Low maximum values for steelhead tended to be early or late in the runs.

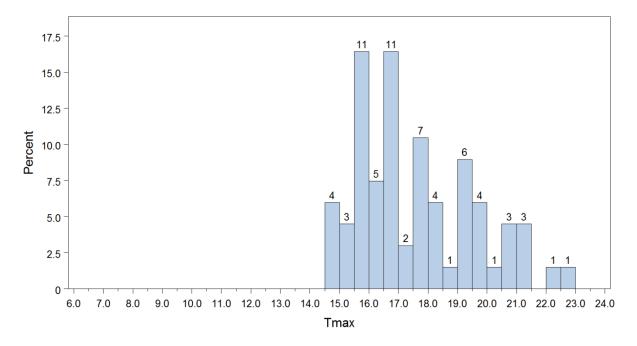


Figure 65. Maximum body temperature of RDST-tagged summer Chinook salmon in The Dalles pool reach, 2000 and 2002. Number above bar = n.

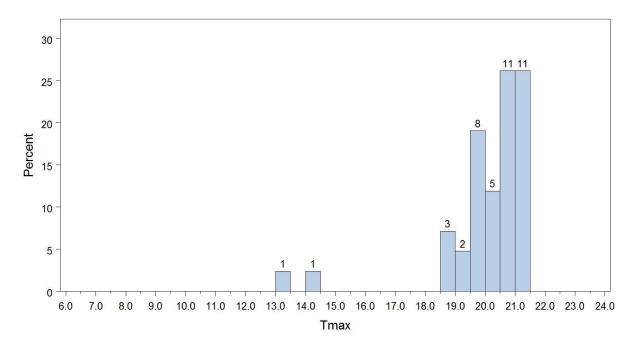


Figure 66. Maximum body temperature of RDST-tagged fall Chinook salmon in The Dalles pool reach, 2000 and 2002. Number above bar = n.

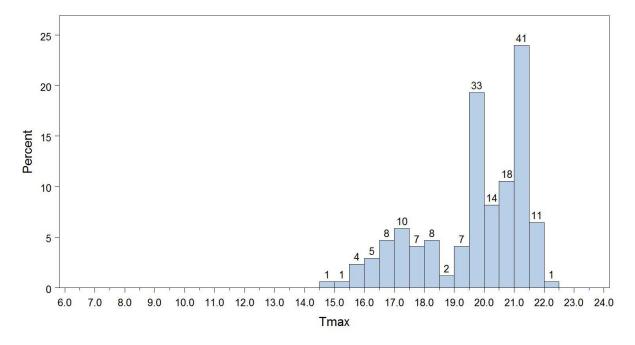


Figure 67. Maximum body temperature of RDST-tagged steelhead in The Dalles pool reach, 2000 and 2002. Number above bar = n.

3.10.3 John Day reservoir reach

Maximum temperatures of summer Chinook salmon that passed McNary Dam ranged from 15.6 °C to 22.2 °C (*mean of maximums* = 18.6 °C, *median of maximums* = 18.4 °C) (Figure 68). Means for fall Chinook salmon ranged from 17.1 °C to 21.7 °C (*mean of maximums* = 20.4 °C, *median of maximums* = 20.7 °C) (Figure 69). Maximums for steelhead ranged from 11.0 °C to 23.0 °C (*mean of maximums* = 19.5 °C, *median of maximums* = 20.0 °C) (Figure 70). Low maximum values for steelhead were mostly from late in the runs.

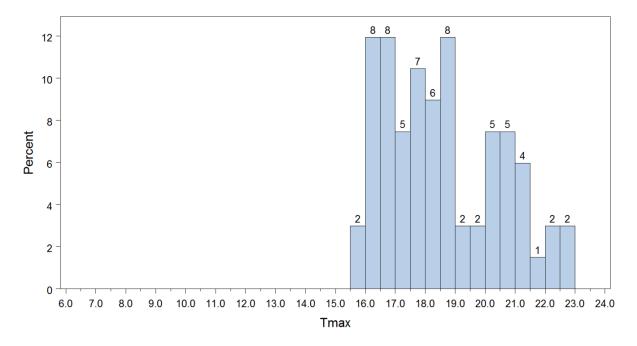


Figure 68. Maximum body temperature of RDST-tagged summer Chinook salmon in the John Day pool reach, 2000 and 2002. Number above bar = n.

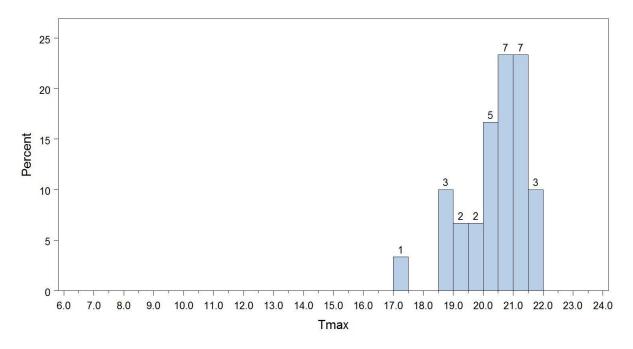


Figure 69. Maximum body temperature of RDST-tagged fall Chinook salmon in the John Day pool reach, 2000 and 2002. Number above bar = n.

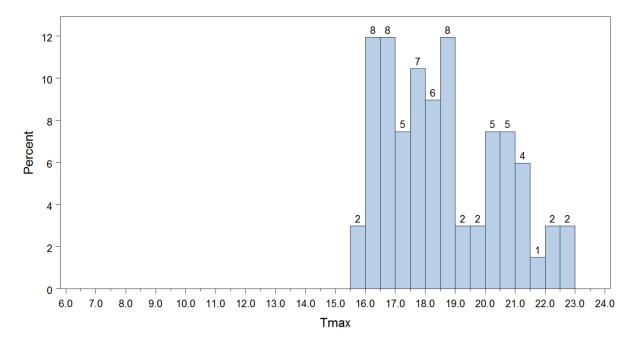


Figure 70. Maximum body temperature of RDST-tagged steelhead in the John Day pool reach, 2000 and 2002. Number above bar = n.

3.10.4 McNary reservoir reach

Maximum temperatures of summer Chinook salmon that passed the McNary pool reach ranged from 14.7 °C to 22.6 °C (*mean of maximums* = 18.4 °C, *median of maximums* = 18.3 °C) (Figure 71). Means for fall Chinook salmon ranged from 12.3 °C to 20.7 °C (*mean of maximums* = 19.5 °C, *median of maximums* = 20.0 °C) (Figure 72). The very cool fall Chinook salmon was a November migrant. Maximums for steelhead ranged from 13.5 °C to 22.2 °C (*mean of maximums* = 19.2 °C, *median of maximums* = 19.3 °C) (Figure 73). Lower maximum values for steelhead were mostly from late in the runs.

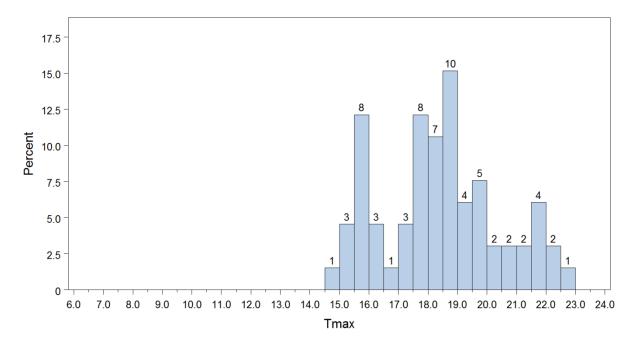


Figure 71. Maximum body temperature of RDST-tagged summer Chinook salmon in the McNary pool reach, 2000 and 2002. Number above bar = n.

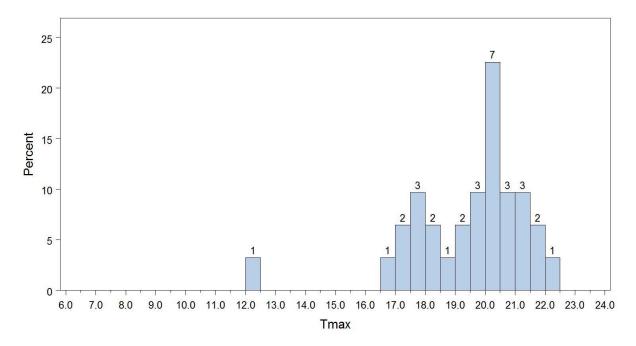


Figure 72. Maximum body temperature of RDST-tagged fall Chinook salmon in the McNary pool reach, 2000 and 2002. Number above bar = n.

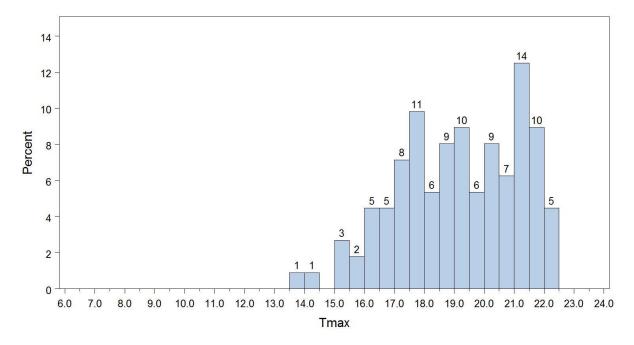


Figure 73. Maximum body temperature of RDST-tagged steelhead in the McNary pool reach, 2000 and 2002. Number above bar = n.

3.11 Cumulative temperature exposure per reach

The degree day (DD) is one of the common metrics used to measure cumulative temperature exposure. We calculated DD (°C above 0) for the RDST-tagged fish in each reservoir reach. These estimates included all times and locations in each reach, including time spent in CWR sites. Fish that were harvested in the reach or that were last detected in tributaries in the reach were excluded from calculations. Some likely underestimation occurred due to tags that reached data storage capacity while fish were in a reach. Distributions of total degree day accumulation may be a useful target metric for HexSim model outputs.

3.11.1 Bonneville reservoir reach

Total DD accumulations for summer Chinook salmon ranged from 13-125 DD (*mean* = 24.0, *median* = 20.1 DD) (Figure 74). Totals for fall Chinook salmon ranged from 16–255 DD (*mean* = 43.4, *median* = 30.2 DD) (Figure 75). Totals for steelhead were much higher and more variable, ranging from 18–755 DD (*mean* = 209.4, *median* = 104.0 DD) (Figure 76).

Of the fish that encountered water temperature ≥ 20 °C, total exposure duration above that threshold averaged 0.96–1.48 d per fish and medians were 0.71–1.15 d per fish (Table 13). Totals included all records in the reach, regardless of location or timing (i.e., they included non-continuous records).

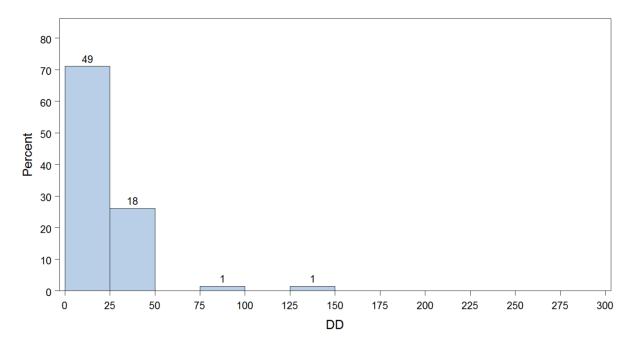


Figure 74. Total accumulated degree days (DD) of RDST-tagged summer Chinook salmon in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

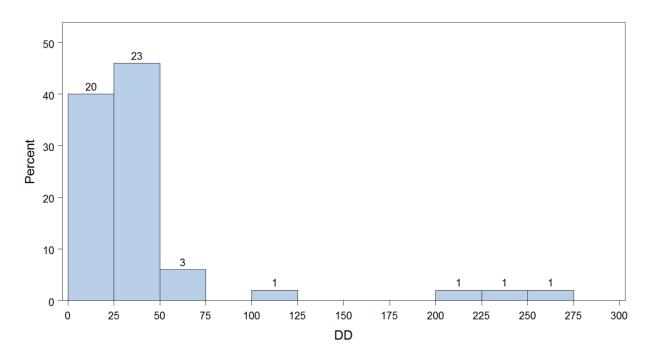


Figure 75. Total accumulated degree days (DD) of RDST-tagged fall Chinook salmon in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

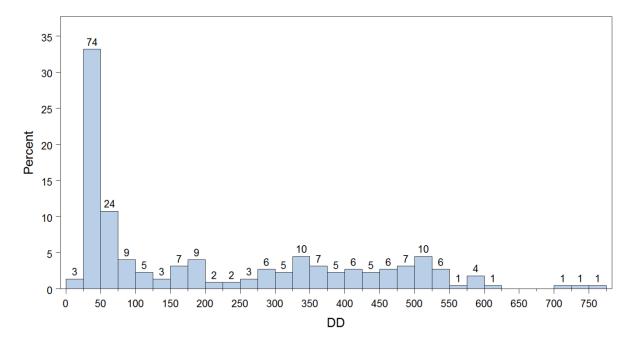


Figure 76. Total accumulated degree days (DD) of RDST-tagged steelhead in the Bonneville pool reach, 2000 and 2002. Number above bar = n.

		$Days \ge 20^{\circ}C$						
Run	Year	$n \geq 20^{\circ} \text{C}$	$\% \ge 20^{\circ}C$	Mean	Median	5%	95%	
Summer Chinook	2000	10	26%	1.25	0.93	0.85	2.23	
	2002	0	0%	-	-	-	-	
Fall Chinook	2000	22	61%	1.48	1.08	0.35	1.99	
	2002	16	73%	1.19	1.02	0.42	2.40	
Steelhead	2000	113	68%	0.96	0.71	0.06	2.22	
	2002	63	55%	1.39	1.15	0.24	2.60	

Table 13. RDST-tagged salmon and steelhead exposure (days) to water temperatures that were ≥ 20 °C while in the Bonneville pool reach.

3.11.2 The Dalles reservoir reach

Total DD accumulations for summer Chinook salmon ranged from 7–266 DD (*mean* = 22.7, *median* = 13.6 DD) (Figure 77). Totals for fall Chinook salmon ranged from 10–118 DD (*mean* = 25.8, *median* = 16.5 DD) (Figure 78). Totals for steelhead were much higher and more variable, ranging from 11–717 DD (*mean* = 131.2, *median* = 34.3 DD) (Figure 79). Of the fish that encountered water temperature ≥ 20 °C, total exposure duration above that threshold averaged 0.48–2.62 d per fish and medians were 0.48–1.46 d per fish (Table 14).

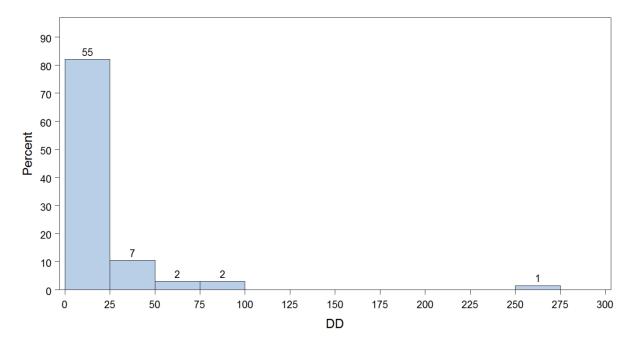


Figure 77. Total accumulated degree days (DD) of RDST-tagged summer Chinook salmon in The Dalles pool reach, 2000 and 2002. Number above bar = n.

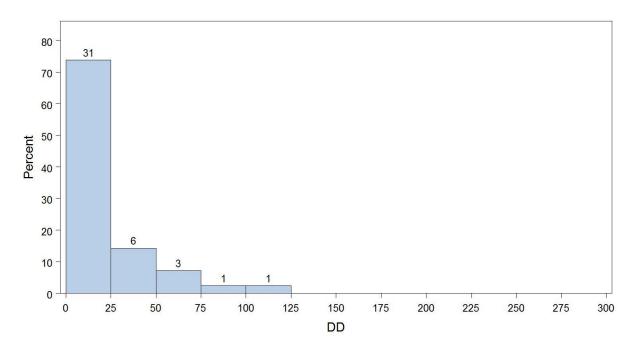


Figure 78. Total accumulated degree days (DD) of RDST-tagged fall Chinook salmon in The Dalles pool reach, 2000 and 2002. Number above bar = n.

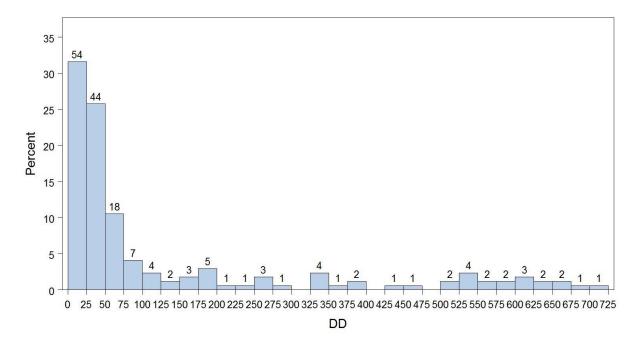


Figure 79. Total accumulated degree days (DD) of RDST-tagged steelhead in The Dalles pool reach, 2000 and 2002. Number above bar = n.

		$Days \ge 20^{\circ}C$						
Run	Year	$n \geq 20^{\circ} \text{C}$	$\% \ge 20^{\circ}C$	Mean	Median	5%	95%	
Summer Chinook	2000	8	22%	1.57	0.76	0.47	4.26	
	2002	1	3%	0.48	0.48	0.48	0.48	
Fall Chinook	2000	16	57%	0.71	0.66	0.48	1.18	
	2002	14	74%	0.80	0.70	0.22	1.71	
Steelhead	2000	27	42%	1.63	0.85	0.08	7.13	
	2002	63	55%	2.62	1.46	0.24	7.28	

Table 14. RDST-tagged salmon and steelhead exposure (days) to water temperatures that were ≥ 20 °C while in the The Dalles pool reach.

3.11.3 John Day reservoir reach

Total DD accumulations for summer Chinook salmon ranged from 19–644 DD (*mean* = 40.0, *median* = 30.0 DD) (Figure 80). Totals for fall Chinook salmon ranged from 28–71 DD (*mean* = 40.7, *median* = 37.2 DD) (Figure 81). Totals for steelhead ranged from 15–195 DD (*mean* = 57.8, *median* = 50.0 DD) (Figure 82). Of the fish that encountered water temperature ≥ 20 °C, total exposure duration above that threshold averaged 0.92–1.72 d per fish and medians were 0.44–1.74 d per fish (Table 15).

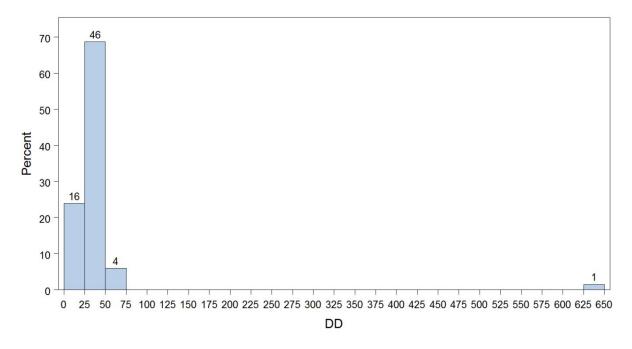


Figure 80. Total accumulated degree days (DD) of RDST-tagged summer Chinook salmon in the John Day pool reach, 2000 and 2002. Number above bar = n.

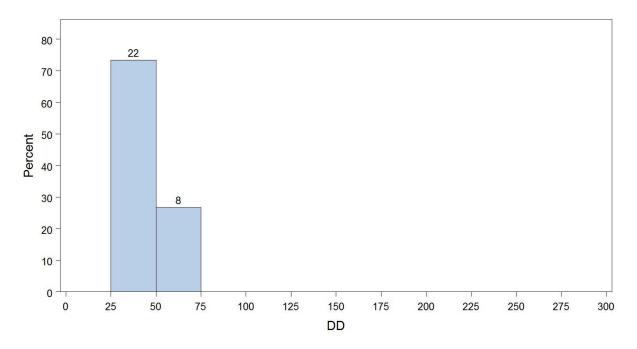


Figure 81. Total accumulated degree days (DD) of RDST-tagged fall Chinook salmon in the John Day pool reach, 2000 and 2002. Number above bar = n.

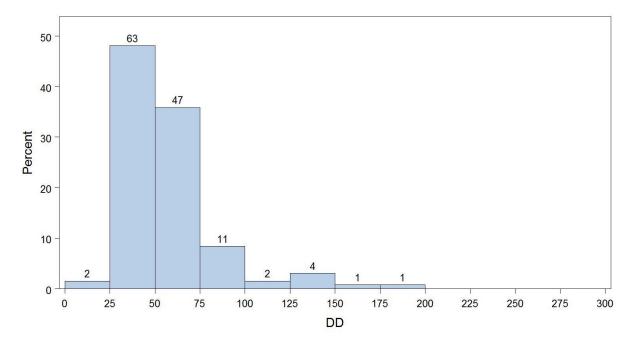


Figure 82. Total accumulated degree days (DD) of RDST-tagged steelhead in the John Day pool reach, 2000 and 2002. Number above bar = n.

		$Days \ge 20^{\circ}C$							
Run	Year	$n \geq 20^{\circ} \text{C}$	$\% \ge 20^{\circ}C$	Mean	Median	5%	95%		
Summer Chinook	2000	0	0%	1.01	1.35	0.02	1.87		
	2002	5	17%	1.45	0.44	0.03	4.67		
Fall Chinook	2000	16	73%	1.42	1.48	0.13	2.90		
	2002	13	72%	0.92	0.94	0.06	1.46		
Steelhead	2000	26	50%	1.45	1.19	0.05	3.06		
	2002	46	53%	1.72	1.74	0.03	3.28		

Table 15. RDST-tagged salmon and steelhead exposure (days) to water temperatures that were ≥ 20 °C while in the John Day pool reach.

3.11.4 McNary reservoir reach

Total DD accumulations for summer Chinook salmon ranged from 6–103 DD (*mean* = 22.0, *median* = 18.3 DD) (Figure 83). Totals for fall Chinook salmon ranged from 15–142 DD (*mean* = 33.3, *median* = 21.5 DD) (Figure 84). Totals for steelhead were more variable and ranged from 17–628 DD (*mean* = 67.0, *median* = 42.5 DD) (Figure 85). Of the fish that encountered water temperature \geq 20 °C, total exposure duration above that threshold averaged 0.36–3.10 d per fish and medians were 0.06–1.74 d per fish (Table 16).

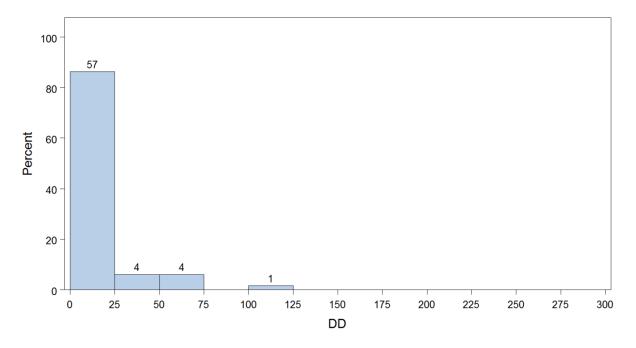


Figure 83. Total accumulated degree days (DD) of RDST-tagged summer Chinook salmon in the McNary pool reach, 2000 and 2002. Number above bar = n.

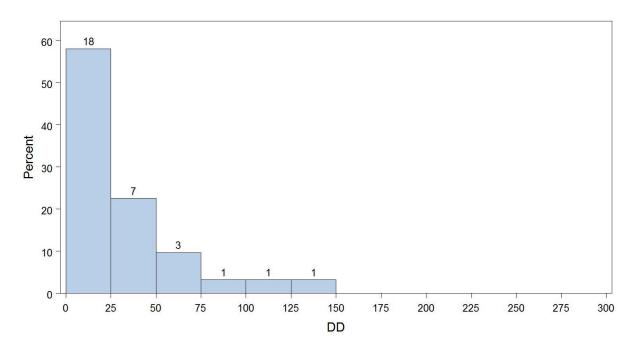


Figure 84. Total accumulated degree days (DD) of RDST-tagged fall Chinook salmon in the McNary pool reach, 2000 and 2002. Number above bar = n.

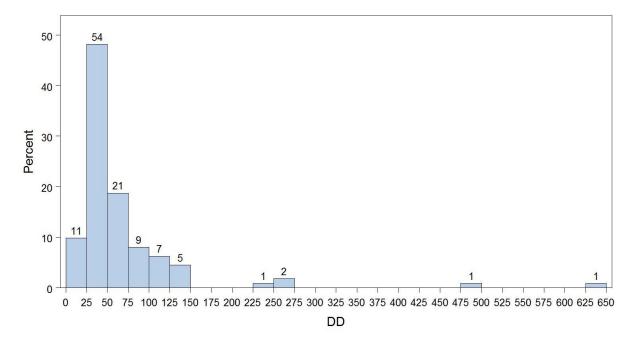


Figure 85. Total accumulated degree days (DD) of RDST-tagged steelhead in the McNary pool reach, 2000 and 2002. Number above bar = n.

		$Days \ge 20^{\circ}C$						
Run	Year	$n \ge 20^{\circ} \text{C}$	$\% \ge 20^{\circ}C$	Mean	Median	5%	95%	
Summer Chinook	2000	10	27%	0.51	0.43	0.27	0.97	
	2002	3	10%	0.39	0.23	0.12	0.77	
Fall Chinook	2000	7	44%	0.89	0.96	0.18	1.51	
	2002	9	60%	0.36	0.06	0.02	1.31	
Steelhead	2000	16	36%	2.11	1.74	0.17	5.17	
	2002	32	43%	3.10	1.38	0.06	13.47	

Table 16. RDST-tagged salmon and steelhead exposure (days) to water temperatures that were ≥ 20 °C while in the McNary pool reach.

3.12 Degree days per day per reach

The scatterplots in this section combine data from previous summaries to illustrate the relationship between cumulative temperature exposure and the time that RDST-tagged fish spent in each reservoir reach. Degree days/day is the rate at which temperature exposure is accumulated, standardized over time. The product (degree days/day) × (days in a reach) is the total degree days presented in Section 3.11.

In general, the warmest DD/d estimates were for salmon and steelhead that moved rapidly through reservoir reaches with little or no CWR use. Relatively cooler DD/d estimates were strongly associated with fish that used CWR sites for multiple days. Some early summer and late fall migrants also had relatively low DD/d.

3.12.1 Bonneville reservoir reach

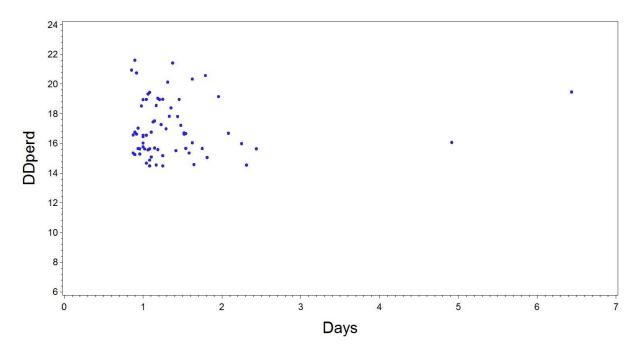


Figure 86. Relationship between the number of days RDST-tagged summer Chinook salmon spent in the Bonneville pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

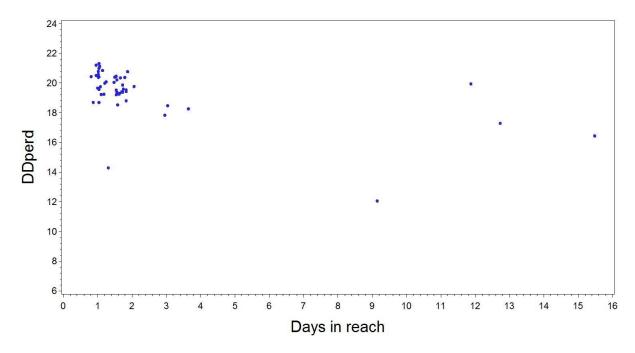


Figure 87. Relationship between the number of days RDST-tagged fall Chinook salmon spent in the Bonneville pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

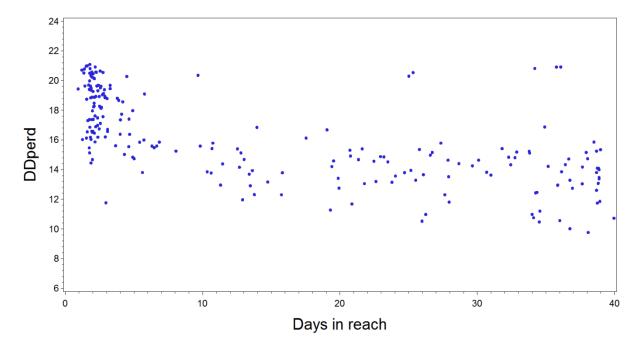
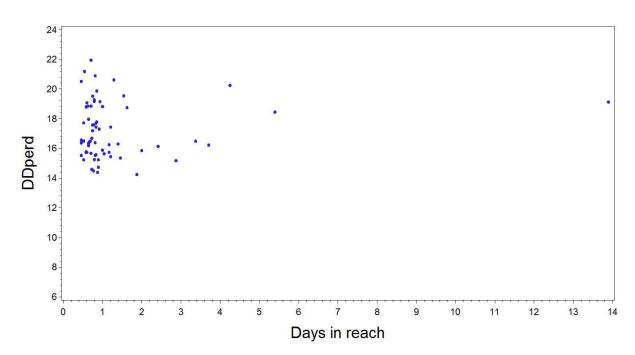


Figure 88. Relationship between the number of days RDST-tagged steelhead spent in the Bonneville pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.



3.12.2 The Dalles reservoir reach

Figure 89. Relationship between the number of days RDST-tagged summer Chinook salmon spent in The Dalles pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

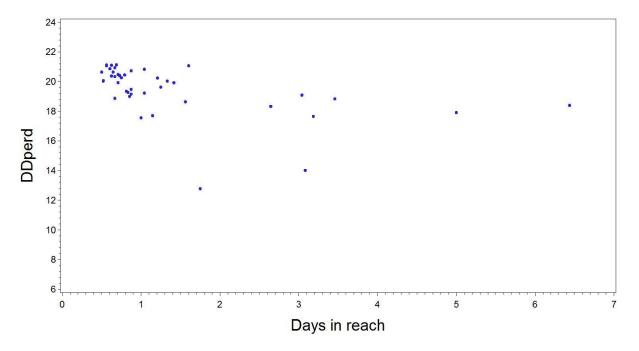


Figure 90. Relationship between the number of days RDST-tagged fall Chinook salmon spent in The Dalles pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

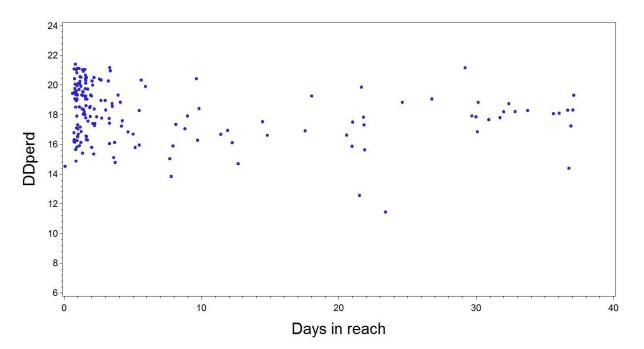


Figure 91. Relationship between the number of days RDST-tagged steelhead spent in The Dalles pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

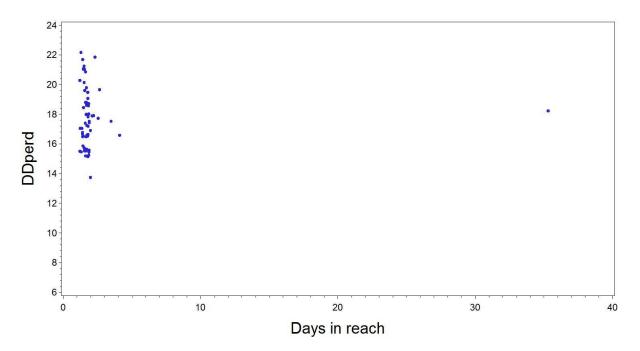


Figure 92. Relationship between the number of days RDST-tagged summer Chinook salmon spent in the John Day pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

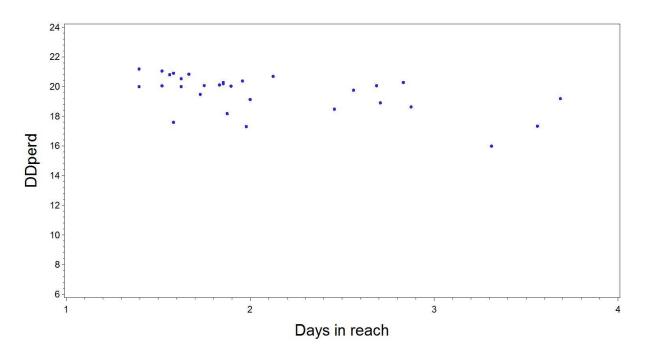


Figure 93. Relationship between the number of days RDST-tagged fall Chinook salmon spent in the John Day pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

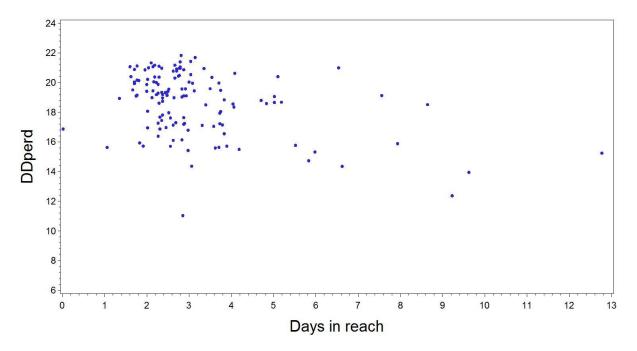
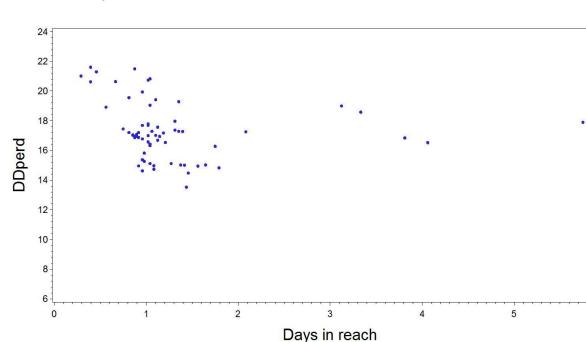


Figure 94. Relationship between the number of days RDST-tagged steelhead spent in the John Day pool reach and the accumulation of degree days per d (DDperd), 2000 and 2002.



3.12.4 McNary reservoir reach

Figure 95. Relationship between the number of days RDST-tagged summer Chinook salmon spent in the McNary reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

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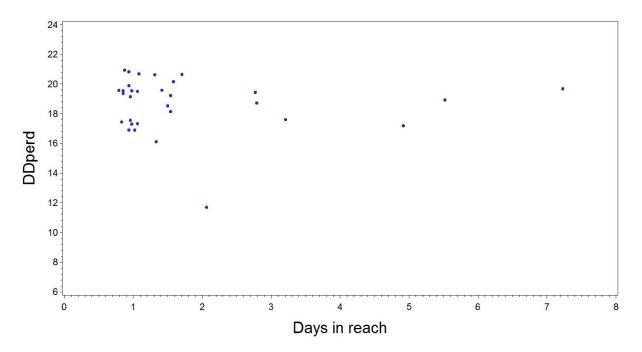


Figure 96. Relationship between the number of days RDST-tagged fall Chinook salmon spent in the McNary reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

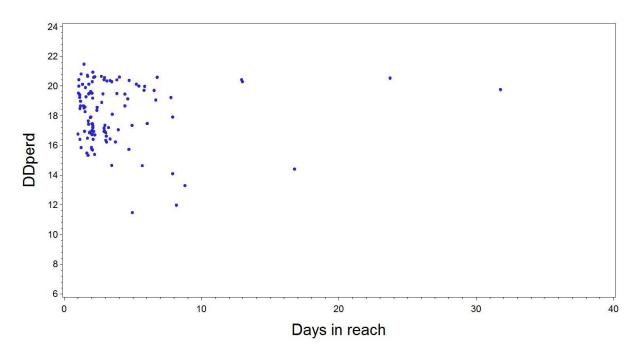


Figure 97. Relationship between the number of days RDST-tagged steelhead spent in the McNary reach and the accumulation of degree days per d (DDperd), 2000 and 2002.

3.13 Probability of CWR use in Bonneville pool by date and temperature

We used logistic regression in exploratory analyses of the relationship between migration date, temperatures RDST-tagged fish encountered in Bonneville reservoir, and use of CWR sites. Fish used in the analyses were restricted to those that passed The Dalles Dam, to reduce ambiguity associated with fish last detected in CWR sites or harvested in the Bonneville pool. Because CWR use by summer Chinook salmon was very limited in this reach, no model was developed. The figures in this section do not substitute for formal analyses, which were beyond the scope of the report, but should provide an indication of overall patterns.

3.13.1 Fall Chinook salmon

Fifty fall Chinook salmon passed The Dalles Dam with suitable RDST data and 12 of these had evidence of CWR use of at least 4 h in Bonneville pool. The relationships between CWR use and both first and maximum encountered temperatures in Bonneville pool indicated weak, slightly decreasing likelihood of CWR use as temperatures increased (Figures 98 and 99). However, results were strongly influenced by a pair of late-run salmon that did use CWR sites. When reservoir entry date was used as the predictor, the predicted probability of use was parabolic-shaped, with relatively high probabilities early and late in migration (Figure 100). The two late-run fish again had relatively high influence on the shape of the plot. We note that the below-average fall temperatures in 2000 and 2002 likely contributed to the low incidence of CWR use by fall Chinook salmon. We recommend reviewing the fall Chinook salmon summaries provided in Goniea et al. (2006), which included more representative samples from a range of water years.

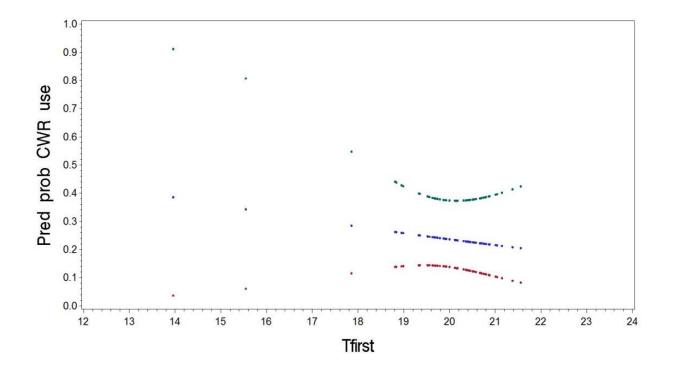


Figure 98. Predicted probability (blue dots) with 95% confidence intervals (red and green dots) that fall Chinook salmon would use a CWR site in Bonneville pool in relation to the first temperature (Tfirst) encountered when salmon entered the pool.

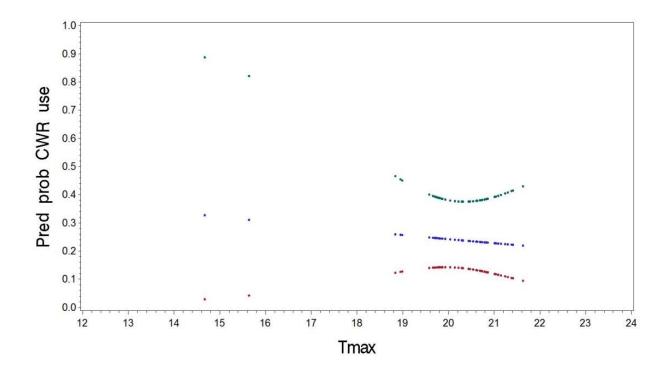


Figure 99. Predicted probability (blue dots) with 95% confidence intervals (red and green dots) that fall Chinook salmon would use a CWR site in Bonneville pool in relation to the maximum temperature (Tfirst) encountered when salmon entered the pool.

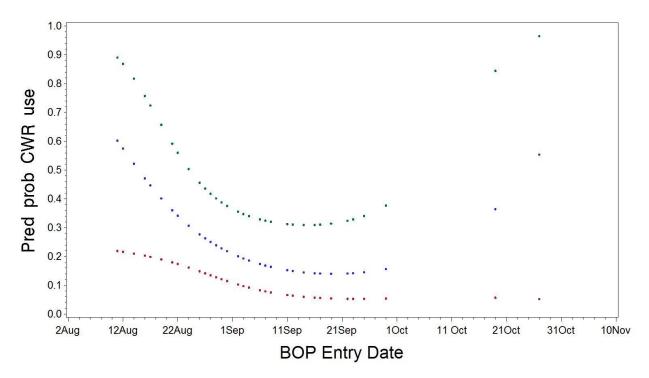


Figure 100. Predicted probability (blue dots) with 95% confidence intervals (red and green dots) that fall Chinook salmon would use a CWR site in Bonneville pool in relation to the date (BOP Entry Date) when salmon entered the pool. The logistic model included a quadratic term for date.

3.13.2 Steelhead

A total of 220 steelhead passed The Dalles Dam with suitable RDST data and 140 of these had evidence of CWR use of at least 4 h in Bonneville pool. The relationships between CWR use and both first increasing likelihood of CWR use as temperatures increased (Figures 101 and 102). When reservoir entry date was used as the predictor, the predicted probability of use was parabolic-shaped, with much higher probabilities from late July through mid-September (Figure 103).

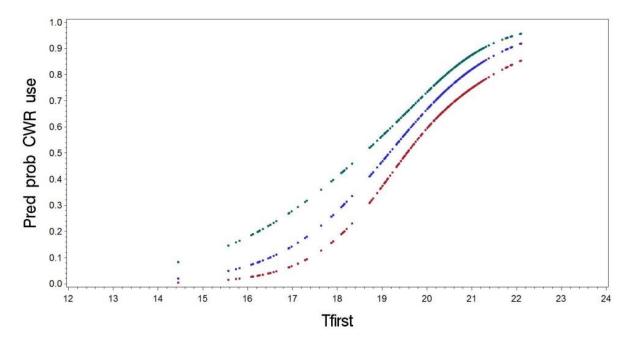


Figure 101. . Predicted probability (blue dots) with 95% confidence intervals (red and green dots) that steelhead would use a CWR site in Bonneville pool in relation to the first temperature (Tfirst) encountered when salmon entered the pool.

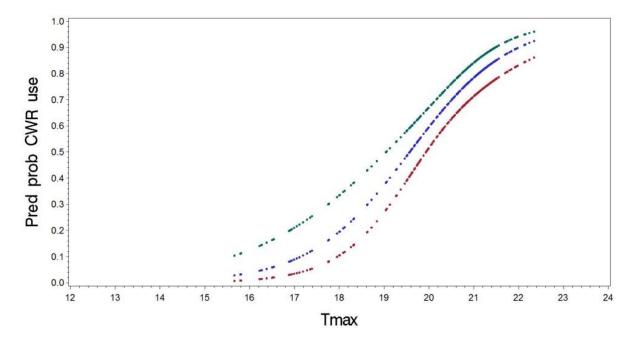


Figure 102. Predicted probability (blue dots) with 95% confidence intervals (red and green dots) that steelhead would use a CWR site in Bonneville pool in relation to the maximum temperature (Tfirst) encountered when salmon entered the pool.

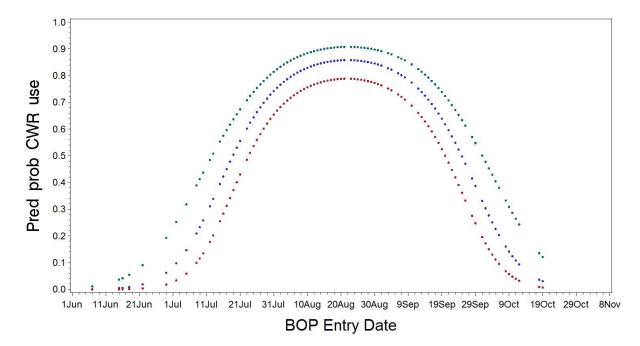


Figure 103. Predicted probability (blue dots) with 95% confidence intervals (red and green dots) that steelhead would use a CWR site in Bonneville pool in relation to the date (BOP Entry Date) when salmon entered the pool. The logistic model included a quadratic term for date.

4.0 References

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